

## Natural Flood Management - Beyond the evidence debate

Journal:	<i>Area</i>
Manuscript ID	AREA-RP-Mar-2018-0032.R2
Manuscript Type:	Regular Paper
Keywords:	Natural Flood Management (NFM), system resilience, catchment partnerships, delivery, catchment management, England and Wales
Abstract:	<p>Globally flood frequency has increased over the last three decades. Natural Flood Management (NFM) is considered a progressive holistic flood management approach, using 'natural' hydrological processes to slow and store water, delivering multiple benefits including water quality, biodiversity and amenity improvements. Although there are existing evaluations of NFM, they remain insufficient for drawing conclusions as to its effectiveness at catchment scales. However, without this evidence base and because of the domination of the natural sciences in the framing and research agenda catchment wide interventions have not been implemented. In acknowledging the importance of understanding and data gaps (and attempts to fill them), this paper argues that there is an opportunity to deliver NFM more widely by capitalising on widespread interest in different land and water management sectors, supported by interdisciplinary policy relevant research. This paper illustrates how multi-stakeholder collaborative partnership is suited to the dynamic complexity of NFM delivery. It is proposed that, through championing NFM delivery at catchment scales and the work of established catchment partnerships in England and Wales, there is the opportunity to more widely deliver NFM as an integrated component of flood risk management.</p>

## 1 Natural Flood Management: An introduction

2 Globally flood frequency has increased over the last three decades (Kundzewicz et al., 2014). Annual  
3 flood losses in the European Union are projected to be c. €23.5 billion by 2050, increased from €4.9  
4 billion between 2000-2012 (Jongman et al., 2014). Traditional flood protection methods of large  
5 engineered structures are costly to construct, expensive to maintain and linked to poor water and  
6 ecological quality (Collentine & Futter, 2016; Thorne, 2014). Policy makers and scientists are  
7 increasingly interested in natural landscape retention and storage capacity as a complementary flood  
8 risk technique (Ran & Nedovic-Budic, 2016) used alongside traditional hard engineering flood defence  
9 approaches (Pitt, 2008; Rouillard et al., 2013; Waylen et al., 2017). However, this move towards more  
10 holistic catchment flood management and natural flood water retention has been hindered by a lack  
11 of policy relevant research with a diversity of disciplinary inputs from both the physical and human  
12 geography traditions. Given the complexity of working at reach to landscape scales and across water  
13 and land management sectors, effective decision making requires an integrated approach, rather than  
14 fragmentation and compartmentalisation, as such the Geography discipline is well situated to support  
15 and develop more holistic approaches to catchment flood management. Several high profile Natural  
16 Flood Management (NFM) projects have been constructed in the UK, such as Belford (Barber & Quinn,  
17 2012), Pickering (Nisbet et al., 2011) and Holnicote (National Trust, 2015). Nevertheless, despite over  
18 a decade of government policy support (Barlow et al., 2014; DEFRA, 2004; Environment Food and  
19 Rural Affairs Committee, 2016; Evans et al., 2004), NFM has still not been a widely adopted as a flood  
20 risk management strategy.

21 NFM harnesses *natural* hydrological processes to slow water flowing through the landscape, thereby  
22 mimicking natural environmental conditions, aspects often lost within traditional flood risk  
23 management (FRM) paradigms of moving water rapidly through the system (Werritty, 2006). The  
24 technical challenge involves an awareness of erosion, sediment transfer and water storage alongside  
25 managing biological inputs, such as, plant growth and nutrient cycling in multiple locations within  
26 catchments (Beechie et al., 2010; Hooke, 2015). NFM is linked to improved biodiversity (Cook et al.,  
27 2016), water quality (Barber & Quinn, 2012; Howe & White, 2003) and public health and well-being  
28 (de Bell et al., 2017; Maas, 2006; Postnote, 2016); a feature referred to within regulatory and  
29 practitioner communities as delivering multiple benefits (Barlow et al., 2014; Forbes et al., 2015).

30 An agreed model of defining principles concerning NFM does not exist, either from a European or UK  
31 perspective. Disagreement on the essential properties of NFM can be found in legislative positioning.  
32 In contrast to England and Wales, the concept of NFM in Scotland is enshrined in law in the Flood Risk  
33 Management (Scotland) Act, Section 20(1) (4) b, with delivery focused on natural features and their

1  
2  
3 34 contribution to FRM. In England, 36% of water bodies are classed as heavily modified, compared to  
4 35 14% in Scotland (FOI request SEPA June 2017), resulting in a greater emphasis on incorporating  
5 36 engineered structures into NFM delivery. 'Working with Natural Processes' (WWNP) has become the  
6 37 favoured term of regulatory authorities in England (Barlow et al., 2014), but lacks clarity resulting from  
7 38 its interchangeable use with NFM (Burgess-Gamble et al., 2017). WWNP is described as emulating  
8 39 natural regulatory functions and applies to any method harnessing hydrological processes within the  
9 40 water cycle, including engineered land forms. WWNP or NFM employs a wide range of techniques  
10 41 applied in diverse settings and across different catchments types, which vary both spatially and  
11 42 temporally (Figure 1). In this paper the broader definition will be used and referred to as NFM. The  
12 43 action of working with the water cycle to increase resilience of the system is taken as the foundation  
13 44 of the NFM paradigm.  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24

25 46 **Figure 1. Catchment wide NFM interventions categorised as the initial step in the hydrological cycle**  
26 47 **– Interception:** A1 bunded ditches, A2 vegetative cover, A3 green roofs and walls, A4 interception  
27 48 ponds, A5 managed realignment, A6 rain gardens, A7 restoring peatlands, A8 swales, A9 beach  
28 49 nourishment, A10 habitat promotion, A11 reef creation **Infiltration:** B1 woodlands , B2 filter/buffer  
29 50 strips, B3 hedgerows, B4 managing soil quality, B5 no and low till agriculture, B6 permeable paving,  
30 51 B7 reduced stocking density **Water storage:** C1 ponds, C2 rainwater harvesting, C3 reservoirs, C4  
31 52 wetlands and reed beds **Channel flow:** D1 de-culverting, D2 increase channel roughness, D3 regulated  
32 53 washlands, D4 remeandering, D5 restore functioning floodplain, D6 setting back flood defences, D7  
33 54 woody material dams D8 species reintroduction (e.g. beavers). Each intervention uses a number of  
34 55 hydrological processes to slow the flow of water for example interception, infiltration and water  
35 56 storage in wetlands and surrounding vegetative cover will result in reduced surface run-off.  
36  
37  
38  
39  
40  
41  
42  
43

## 44 57 **Evidence**

45  
46  
47 58 One of the factors purported to be behind limited adoption of NFM is lack of evidence of its efficacy  
48 59 (Dadson et al., 2017; Spray et al., 2009; Waylen et al., 2017). Dadson *et al.* (2016) cite a lack of  
49 60 demonstrable effects beyond small-scale local benefits, observing that an absence of evidence exists  
50 61 of interventions tested at large-scales. Difficulties in translating findings between sites/catchments  
51 62 are compounded by the inherent low frequency of flood events, with each flood characterised by a  
52 63 number of variables. As such, gathering sufficient data to draw statistically robust conclusions at  
53 64 catchment scales may take decades, or be unachievable (Pattison & Lane, 2012).  
54  
55  
56  
57  
58  
59  
60

1  
2  
3 65 FRM investment is informed by risk based predictive modelling (Woodward et al., 2009), that  
4 66 quantifies current and future risk (Environment Agency, 2016b). Multiple sources of flooding – fluvial,  
5 67 pluvial, coastal and groundwater – are often analysed in isolation when, in reality, they act in  
6 68 combination. Hydrological response is determined by spatial processes e.g. land use, geology,  
7 69 vegetation, urban drainage capacity and structures and temporal processes of antecedent conditions  
8 70 and rainfall event sequencing (Miller et al., 2014). The physical processes a model represents -  
9 71 whether statistical, conceptual or physically based – are simplified system representations, requiring  
10 72 at times, ill-fitting simplified generalised parameters (Redfern et al., 2016). The level of complexity e.g.  
11 73 data requirements, computational power, model parameters and functions, rapidly escalates as  
12 74 catchment features and interactions are combined to create increasingly realistic replications of  
13 75 catchment process and responses (Mcintyre & Thorne, 2013; Metcalfe et al., 2017). Therefore,  
14 76 catchment size becomes important. As NFM is an *integrated* system of local interventions performed  
15 77 at varying scales, predictive methods need to move between field and catchment scales. The data  
16 78 inputs required when mapping or modelling catchment processes are influenced by required output  
17 79 scales, but also consideration of the efficiency of integrating spatial data (Medcalf, *et al.* 2014).  
18 80 Methodologically, it is difficult to demonstrate any single individual NFM interventions contribution  
19 81 to flood risk, as hydrograph changes may be hard to detect at short temporal or above local scales  
20 82 (Metcalfe et al., 2017). This makes separating causal factors from natural or human induced catchment  
21 83 or meteorological changes challenging.

22 84 The influence of computational limitations to modelling are creeping into NFM practitioner guidance;  
23 85 demonstrated with the recommendation that suitability is limited to small catchments in rural  
24 86 headwaters of  $\leq 10\text{km}^2$  (Environment Agency, 2016a; National Trust, 2015; Westcountry Rivers Trust,  
25 87 2014). The overriding factor leading to a focus on small catchments alone comes from a need to  
26 88 demonstrate a measurable benefit to flood risk reduction through, monitoring, hydrological modelling  
27 89 or a cost-benefit analysis. As such, guidance is not leading practitioners to understand the catchment  
28 90 *as a system*, or to increase resilience by building capacity into the landscape and maintaining desirable  
29 91 hydrological regimes (Liao, 2012).

30 92 The risk, in delaying adoption of catchment wide NFM until 'sufficient' empirical evidence exists means  
31 93 the full potential of NFM will not be realised and its application as a technique in FRM will remain '*ad*  
32 94 *hoc*'. This paper argues - given longstanding interest from policy makers and practitioners, and  
33 95 consensus around delivery of multiple benefits - research and resources should be expanded beyond  
34 96 a principles, evidence and efficacy debate to mechanisms of NFM delivery. The natural sciences  
35 97 dominate NFM research literature, with limited social science analysis. It is against this backdrop, of  
36 98 translating policy into practice, that a review of sectors relevant to NFM delivery is undertaken with

1  
2  
3 99 the aim of understanding their roles and interactions and the strength and weaknesses of associated  
4  
5 100 policy objectives. Rather than viewing NFM as a unique environmental management challenge,  
6  
7 101 synergies with other water management frameworks have been investigated, particularly those with  
8  
9 102 links to relevant sectors, with the aim of identifying existing forums that can be utilised to encourage  
10  
11 103 wider delivery.

12 104

## 15 105 **Who should contribute to the delivery of NFM?**

17 106 UK water management and regulation is divided amongst different organisations into functional  
18  
19 107 elements of water quality, quantity, freshwater ecology, hydro-power and recreational activities,  
20  
21 108 resulting in fragmented regulatory and administration duties (Howarth, 2017). NFM as a flood  
22  
23 109 management technique could be assumed to be an exclusive concern of the Flood and Coastal Risk  
24  
25 110 Management (FCRM) sector, however, as NFM requires multiple interventions across the landscape  
26  
27 111 (Figure 1), a range of land uses and stakeholders are essential in delivery. The chief sectors are: urban  
28  
29 112 planning and development (Ellis, 2013), FCRM (Thorne, 2014), water supply (Kidd & Shaw, 2007),  
30  
31 113 agriculture and nature conservation (Acreman & Holden, 2013; Howe & White, 2003). These sectors  
32  
33 114 need to interact cooperatively, but effective large scale change usually requires collaboration and  
34  
35 115 coordination, the scale of which can be evaluated through shared policy objectives, planning and  
36  
37 116 delivery mechanisms (Robins et al., 2017).

38 117 The principal aspiration of the Department for Communities and Local Government is to increase  
39  
40 118 housing supply (DCLG, 2016), conversely the Environment Agency aims to reduce flood and coastal  
41  
42 119 erosion risk (Environment Agency, 2015). Inevitably this can place the two organisations in opposing  
43  
44 120 perspectives, as development sites are often on floodplains adjacent to watercourses. The continued  
45  
46 121 expansion of floodplain developments to build new homes (CCC, 2012) results in an increasing number  
47  
48 122 of properties at risk, which the Environment Agency must assess for protection. Planning authorities  
49  
50 123 must take the probability of flooding into account when determining applications, therefore all  
51  
52 124 planning applications within a designated flood zone or area with critical drainage problems must also  
53  
54 125 include a flood risk assessment. Nevertheless, construction has grown in areas of high flood risk at a  
55  
56 126 rate of 1.2% since 2011 (CCC, 2015). If the planning system was employed to designate, protect and  
57  
58 127 reinstate functional floodplains fewer homes would require flood protection, reducing downstream  
59  
60 128 flood risk, and communities could enjoy the multiple benefits provided by green spaces.

61 129 The sewer network is probably the most familiar, yet overlooked, rain water management system. In  
62  
63 130 the UK, private water companies manage the majority of the sewer network (Blackburn et al., 2017).

1  
2  
3 131 By employing NFM measures in developments, such as swales, basins, ponds and wetlands, volumes  
4  
5 132 of water entering the sewerage system and its associated costs can be reduced (Stevens & Ogunyoye,  
6  
7 133 2012). In this context NFM techniques are most commonly referred to as SuDS (Sustainable Drainage  
8  
9 134 Systems) (Figure 1: Interventions A3, A4, A6, A8, B6, C2). Water companies are supportive of making  
10  
11 135 changes to the drainage infrastructure necessary to employ SuDS more widely (Lieberherr & Truffer,  
12  
13 136 2015; Water UK, 2015). Initially, the *Flood and Water Management Act 2010* was to drive this change  
14  
15 137 and make SuDS a legal requirement in new developments (Walker et al., 2011). However, the  
16  
17 138 government maintained existing arrangements through the planning system and introduced non-  
18  
19 139 statutory technical standards instead, which removed reference to improved water quality,  
20  
21 140 biodiversity or access to green space. Currently SuDS are an expectation rather than a legal  
22  
23 141 requirement for developments of 10 properties or more (Lewis, 2014). The implementation of  
24  
25 142 Schedule 3 of the *Flood and Water Management Act 2010* can be considered a missed opportunity for  
26  
27 143 SuDS delivery and unlikely to deliver the desired change. A survey in 2014 showed only a quarter of  
28  
29 144 planning applications specified the inclusion of any SuDS measures (CCC, 2014). Rather than regulatory  
30  
31 145 measures encouraging collaborative working, current policy relies on developers to act voluntarily,  
32  
33 146 causing water companies to publicly express discomfort, whilst distancing themselves from  
34  
35 147 implementation (CIWEM., 2017; Water UK, 2015). If SuDS were a statutory requirement in new  
36  
37 148 developments, there would be an opportunity to deliver their multiple benefits (Walker et al., 2011).  
38  
39 149 Furthermore, the adoption of national SuDS standards, would facilitate the incorporation of  
40  
41 150 interventions into existing integrated sewer networks, permitting it to function as one integrated  
42  
43 151 system (Jones & Macdonald, 2007).

39 152 Responding to growing evidence of freshwater habitat degradation and impacts on communities and  
40  
41 153 livelihoods, principles were developed, for a co-ordinated approach to manage water and land as one  
42  
43 154 system, ensuring equitable availability and sustainable use; providing the basis for the international  
44  
45 155 movement of Integrated Water Resource Management (IWRM) (Calder 2005). In Europe IWRM  
46  
47 156 translated into legislation through the *Water Framework Directive* (Directive 2000/60/EC) (WFD) and  
48  
49 157 river basin planning (Rahaman et al., 2004; Richter et al., 2013); followed seven years later by the *EU*  
50  
51 158 *Floods Directive* (FD) (2007/60/EC). The EU Environment Ministers agreed that member states should  
52  
53 159 “maximise the synergies” through integrated approaches to directive delivery (Neuhold, 2014)  
54  
55 160 aligning objectives through the planning instruments of River Basin Management Plans (RBMP) and  
56  
57 161 Flood Risk Management Plans (FRMP); which use the same geographical boundaries and share similar  
58  
59 162 6-year planning cycles (Environment Agency 2015). With clear targets for improving the ecological  
60  
163 status of waterbodies, the delivery of WFD has been the sole focus of RBMP, but the scope to do more  
164  
has always existed (Robins et al., 2017); in contrast FD is procedural. How flood protection is achieved

1  
2  
3 165 is devolved to the member states (Newig et al., 2014). Therefore, the commitment to change to an  
4  
5 166 integrated catchment system approach must come from organisations managing flood risk, rather  
6  
7 167 than guided directly by the directive. The integration of water provisioning organisations and FCRM  
8  
9 168 through the synergies of WFD and FD provides an opportunity to promote NFM and delivery of  
10  
11 169 multiple benefits (Barlow et al., 2014; Collentine & Futter, 2016) but the operational integration of  
12  
13 170 the two directives is not subjected to a formal review within the UK.

14 171 To combat the decline in biodiversity, agri-environment schemes were introduced in 1987 to  
15  
16 172 encourage farming systems that enhance and conserve biodiversity (Batáry et al., 2015; Boatman et  
17  
18 173 al., 2008). Through review and reform, the schemes' initial emphasis on preventing species loss  
19  
20 174 evolved to maintenance and improvement of ecosystem services (Batáry et al., 2015). The agri-  
21  
22 175 environment schemes in England, known as Countryside Stewardship, were introduced in 2016 and  
23  
24 176 includes NFM interventions (Defra et al., 2016), and encouraged landowner/ farmer participation  
25  
26 177 (Riley et al., 2018). However, contributions of agri-environment schemes to delivering catchment wide  
27  
28 178 NFM is being overlooked by the wider FRM community. In recently published FRMPs (Environment  
29  
30 179 Agency, 2016a) only one of seven plans explicitly mentions agri-environment schemes. The majority  
31  
32 180 note greater future engagement with the agricultural sector; however, two plans fail to link farming  
33  
34 181 and flood management at all (Environment Agency, 2016a).

33  
34  
35  
36 182

### 36 183 **Harnessing synergies**

37  
38  
39 184 The shift from the dominance of hard engineered flood defences, designed to defend individual  
40  
41 185 communities, to an integrated system operating at a catchment level, requires an increased number  
42  
43 186 of professionals and organisations working together. As illustrated, cross-sectoral scope does not fit  
44  
45 187 neatly into existing working patterns and governance mechanisms (Rouillard et al., 2013). Accordingly,  
46  
47 188 NFM and the WFD are identical (Collentine & Futter, 2016) and require coordinated action of the same  
48  
49 189 sectors to manage water and land as one system .

49  
50 190 In 2011 the English government piloted an integrated water management initiative, 'The Catchment  
51  
52 191 Based Approach' (CaBA). Following two years of trials, the scheme was extended (Defra, 2013b).  
53  
54 192 Catchment partnerships are led by host charitable organisations, with little guidance from government  
55  
56 193 and have considerable freedom to develop collaborative arrangements focused towards local  
57  
58 194 circumstances and conditions (Watson, 2015). They are encouraged to produce a catchment plan  
59  
60 195 enabling them to implement a range of interventions to realise their goals including strategy planning  
60 196 over a long-term horizon, codes of working, mission statements and project based activities



1  
2  
3 197 (Hurlimann & Wilson, 2018). The Environment Agency administers funding on behalf of Defra, with  
4  
5 198 activities expected to support partnerships to help deliver RBMP objectives (Defra, 2015), foster local  
6  
7 199 collaboration, and deliver multiple benefits (Cascade, 2013; Defra, 2013a). Through mutual  
8  
9 200 recognition of the interdependence each partnership organisation has on a healthy water  
10  
11 201 environment, sufficient motivation should exist to resolve conflicts of interest (Smith, 2015) and  
12  
13 202 deliver more ambitious environmental improvements. CaBA has always been intended to be a  
14  
15 203 mechanism for better integration of FRM into integrated catchment management (Defra, 2013a),  
16  
17 204 couple this with the fact that they already deliver many NFM interventions (figure 1) (CaBA, 2017)  
18  
19 205 motivated by objectives other than FRM. Yet the scale, efficiency, outputs and outcomes of delivery  
20  
21 206 compared to rhetoric are unknown. It is suggested that without statutory reform, and increased  
22  
23 207 funding, catchment partnerships are limited in ability to deliver IWRM with or without FRM (Robins  
24  
25 208 et al., 2017; Watson, 2015). Whilst the collaborative catchment model is not yet proven in England a  
26  
27 209 review of such partnerships, principally in USA and Australia, identified factors that promote  
28  
29 210 successful outcomes; including the necessity of adequate funding, effective leadership, trust, and  
30  
31 211 commitment of partners (Leach & Pelkey, 2001). It is unknown whether CaBA has embedded these  
32  
33 212 lessons into partnerships but it is likely that the success of promoting delivery of WFD and FD through  
34  
35 213 NFM will be heavily influenced by these factors.

36  
37 214 Catchment partnerships, whilst not responsible for WFD delivery, are aligned to the directive as  
38  
39 215 illustrated by 100% of projects undertaken by 25 pilot CaBA catchments focused on water quality.  
40  
41 216 Catchment partnerships often frame collaborative catchment management as an effective means for  
42  
43 217 delivery, with a focus on tackling specific 'local' issues for improving water or habitat quality, rather  
44  
45 218 than a holistic whole systems approach (Watson, 2015). If reflected in the delivery of NFM, CaBA is in  
46  
47 219 a strong position to deliver multiple benefits, however, limited evidence exists that suggests that NFM  
48  
49 220 objectives of preventing, protecting and mitigating flood risk are well understood by catchment  
50  
51 221 partnerships in their current form. Therefore, the recommendation to increase delivery of NFM  
52  
53 222 through catchment partnerships is dependent on strong engagement with the FCRM sector.

54  
55 223 The House of Commons Environment Food and Rural Affairs Committee reviewed current flood risk  
56  
57 224 management delivery in England (Environment Food and Rural Affairs Committee, 2016; Howarth,  
58  
59 225 2017) and concluded that NFM type measures were needed and wider scale adoption through local  
60  
226 stakeholder partnerships encouraged. Catchment partnerships are referenced within the report, but  
227 only as a source of information or as a contractor to deliver projects, illustrating a lack of  
228 understanding of their current purpose and potential in delivering a resilient hydrological system,  
229 overlooking synergies in FD and WFD. No formal UK assessment of whether the delivery of the two  
230 directives have been integrated operationally has been undertaken, a 2017 review of catchment



1  
2  
3 231 partnerships reported 99% contributed to WFD objectives. Whilst a similar figure is not given for  
4  
5 232 contributing to FD, 91% of catchment partnerships report engaging with FRM (CaBA, 2017). It is our  
6  
7 233 proposal that CaBA is well placed to realise the integration of FRM into IWRM in England but is yet to  
8  
9 234 be tested in practice and an area for future research.

10  
11 235

## 12 13 14 236 **Conclusion**

15  
16  
17 237 When the word natural in NFM is taken to refer to hydrological processes rather than related to the  
18  
19 238 idea of the naturalness of a feature, the catchment can be considered as an integrated system. NFM  
20  
21 239 is designed to increase the resilience within the system by creating capacity for slowing and storing  
22  
23 240 water, rather than moving it as quickly as possible. However, over ten years of policy documentation  
24  
25 241 and a widespread interest in NFM-type interventions has not resulted in NFM being adopted as a  
26  
27 242 mainstream FRM strategy. To date NFM is typically implemented on a small scale, *ad hoc*,  
28  
29 243 unsystematic basis, despite the benefits to wildlife and society. The lack of widespread adoption could  
30  
31 244 be due to the focus of research and resources to increase the evidence base; a complex and lengthy  
32  
33 245 process and paradoxically unobtainable if NFM is not applied at the catchment scale. Moreover, rather  
34  
35 246 than embracing the notion of creating a more resilient system, the computational complexities of  
36  
37 247 increasing our knowledge base almost entirely through modelling is leading to a narrowing of the  
38  
39 248 scope of NFM away from a systems approach to small rural catchments.

40  
41 249 Catchment wide NFM collates expertise and responsibilities from a wide range of sectors requiring  
42  
43 250 collaborative working. This review of sectors has revealed policy objectives are not presently aligned,  
44  
45 251 with a divergence in activities rather than coordinated cooperative planning and working. This  
46  
47 252 divergence is further fuelled by a paucity of interdisciplinary relevant policy research, capable of  
48  
49 253 binding together physical observations and projections with long term policy planning and water  
50  
51 254 management frameworks. The unsystematic informal NFM delivery to date could be tied to a lack of  
52  
53 255 research into delivery mechanisms.

54  
55 256 The two philosophies behind CaBA and NFM are comparable and compatible; working at the  
56  
57 257 catchment scale and delivering multiple benefits. Moreover, the contributing sectors are identical.  
58  
59 258 This is a solid base from which to coordinate delivery. Therefore, rather than creating a new  
60  
259 partnership, an opportunity exists to utilise and develop existing synergies between the FD and WFD  
260  
261 and seek to champion NFM through established catchment partnerships. If CaBA is supported through  
research into the mechanisms of delivery, NFM could be realised at the required catchment scale,

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60

262 which would enable wider recognition of NFM and assessment as a mainstream flood risk  
263 management strategy.

264

For Peer Review

265 **References**

- 266 Acreman, M., & Holden, J. (2013). How wetlands affect floods. *Wetlands*, 33(5), 773–786.  
267 <https://doi.org/10.1007/s13157-013-0473-2>
- 268 Barber, N. J., & Quinn, P. F. (2012). Mitigating diffuse water pollution from agriculture using soft-  
269 engineered runoff attenuation features. *Area*, 44(4), 454–462. [https://doi.org/10.1111/j.1475-](https://doi.org/10.1111/j.1475-4762.2012.01118.x)  
270 [4762.2012.01118.x](https://doi.org/10.1111/j.1475-4762.2012.01118.x)
- 271 Barlow, J., Moore, F., & Burgess-Gamble, L. (2014). *Working with natural processes to reduce flood*  
272 *risk: R & D framework*.
- 273 Batáry, P., Dicks, L. V., Kleijn, D., & Sutherland, W. J. (2015). The role of agri-environment schemes in  
274 conservation and environmental management. *Conservation Biology*, 29(4), 1006–1016.  
275 <https://doi.org/10.1111/cobi.12536>
- 276 Beechie, T. J., Sear, D. A., Olden, J. D., Pess, G. R., Buffington, J. M., Moir, H., ... Pollock, M. M. (2010).  
277 Process-based Principles for Restoring River Ecosystems. *BioScience*, 60(3), 209–222.  
278 <https://doi.org/10.1525/bio.2010.60.3.7>
- 279 Blackburn, H., O’Neill, R., & Rangeley-Wilson, C. (2017). Flushed away: how sewage is still polluting  
280 the rivers of England and Wales, 1–37. Retrieved from  
281 [https://www.wwf.org.uk/sites/default/files/2017-12/Flushed Away\\_\\_Nov2017.pdf](https://www.wwf.org.uk/sites/default/files/2017-12/Flushed Away__Nov2017.pdf)
- 282 Boatman, N., Ramwell, C., Parry, H., Bishop, J., Gaskell, P., Mills, J., & Dwyer, J. (2008). A review of  
283 environmental benefits supplied by agri-environment schemes, (August), 275.
- 284 Burgess-Gamble, L., Ngai, R., Wilkinson, M., Nisbet, T., Pontee, N., Harvey, R., ... Quinn, P. (2017).  
285 Working with Natural Processes – Evidence Directory, 311.
- 286 CaBA. (2017). Monitoring & Evaluation and, (November), 6002.
- 287 Cascade. (2013). Defra Evaluation of the Catchment Based Approach - Pilot Stage Final Evaluation  
288 Report, (May).
- 289 CIWEM., and W. (2017). *A place for SuDS? 1*. Retrieved from [http://www.ciwem.org/wp-](http://www.ciwem.org/wp-content/uploads/2017/10/A-Place-for-SuDS-Online.pdf)  
290 [content/uploads/2017/10/A-Place-for-SuDS-Online.pdf](http://www.ciwem.org/wp-content/uploads/2017/10/A-Place-for-SuDS-Online.pdf)
- 291 Collentine, D., & Futter, M. N. (2016). Realising the potential of natural water retention measures in  
292 catchment flood management: trade-offs and matching interests. *Journal of Flood Risk*  
293 *Management*. <https://doi.org/10.1111/jfr3.12269>

- 1  
2  
3 294 Committee on Climate Change. (2012). Climate change - is the UK preparing for flooding and water  
4 scarcity? Adaptation Sub-Committee Progress Report 2012.  
5 295  
6  
7 296 Committee on Climate Change. (2014). Managing climate risks to well-being and the economy, 202.  
8  
9  
10 297 Committee on Climate Change. (2015). Reducing emissions and preparing for climate change: 2015  
11 298 Progress Report to Parliament, (June), 1–36. Retrieved from  
12 papers3://publication/uuid/112E7E63-A7B8-47D6-874E-40D6430B6479  
13 299  
14  
15 300 Cook, B., Forrester, J., Bracken, L., Spray, C., & Oughton, E. (2016). Competing paradigms of flood  
16 management in the Scottish/English borderlands. *Disaster Prevention and Management: An*  
17 301 *International Journal*, 25(3), 314–328. <https://doi.org/10.1108/DPM-01-2016-0010>  
18 302  
19  
20  
21 303 Dadson, S. J., Hall, J. W., Murgatroyd, A., Acreman, M., Bates, P., Beven, K., ... Wilby, R. (2017). A  
22 restatement of the natural science evidence concerning catchment-based “natural” flood  
23 304 management in the UK. Retrieved from  
24 305  
25 306 <http://dx.doi.org/10.1098/rspa.2016.0706> <https://dx.doi.org/10.6084/m9.0Ahttp://rspa.royalsocietypublishing.org/>  
26 307  
27  
28  
29  
30 308 DCLG. (2016). Single departmental plan: 2015 to 2020. Retrieved November 21, 2017, from  
31 [https://www.gov.uk/government/publications/dclg-single-departmental-plan-2015-to-](https://www.gov.uk/government/publications/dclg-single-departmental-plan-2015-to-2020/single-departmental-plan-2015-to-2020)  
32 309 [2020/single-departmental-plan-2015-to-2020](https://www.gov.uk/government/publications/dclg-single-departmental-plan-2015-to-2020/single-departmental-plan-2015-to-2020)  
33 310  
34  
35  
36 311 de Bell, S., Graham, H., Jarvis, S., & White, P. (2017). The importance of nature in mediating social  
37 and psychological benefits associated with visits to freshwater blue space. *Landscape and*  
38 312 *Urban Planning*, 167(May), 118–127. <https://doi.org/10.1016/j.landurbplan.2017.06.003>  
39 313  
40  
41  
42 314 Defra. (2013a). Catchment Based Approach: Improving the quality of our water environment.  
43 *Department for Environment, Food & Rural Affairs*, (May), 1–32. Retrieved from  
44 315 [https://www.gov.uk/government/publications/catchment-based-approach-improving-the-](https://www.gov.uk/government/publications/catchment-based-approach-improving-the-quality-of-our-water-environment)  
45 316 [quality-of-our-water-environment](https://www.gov.uk/government/publications/catchment-based-approach-improving-the-quality-of-our-water-environment)  
46 317  
47  
48  
49 318 Defra. (2013b). Guide to Collaborative Catchment Management, (August), 68. Retrieved from  
50 319 <http://ccmhub.net/wp-content/uploads/2012/10/The-Guide.pdf>  
51  
52  
53 320 DEFRA. (2004). Making space for water: Developing a new Government strategy for flood and  
54 321 coastal erosion risk management in England., (July), 1–156.  
55  
56  
57 322 Defra, EAFRD, & Natural England. (2016). Countryside Stewardship - How can it help reduce  
58 323 flooding ? Retrieved from  
59 324

- 1  
2  
3 325 flood-reduction-infographic.pdf  
4  
5  
6 326 Department for Environment Food and Rural Affairs. (2015). *Catchment Partnership Fund :*  
7 327 *Environment Agency Summary Report 2014-2015 Catchment Partnership Fund : Environment*  
8 *Agency Summary Report 2014-2015*. Retrieved from  
9 328  
10 329 [https://www.gov.uk/government/uploads/system/uploads/attachment\\_data/file/492857/catch-](https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/492857/catchment-partnership-fund-report-2014-2015.pdf)  
11 [ment-partnership-fund-report-2014-2015.pdf](https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/492857/catchment-partnership-fund-report-2014-2015.pdf)  
12 330  
13  
14 331 Ellis, J. B. (2013). Sustainable surface water management and green infrastructure in UK urban  
15 332 catchment planning. *Journal of Environmental Planning and Management*, 56(1), 24–41.  
16 333 <https://doi.org/10.1080/09640568.2011.648752>  
17  
18  
19  
20 334 Environment Agency. (2015). Managing flood and coastal erosion risks in England, (March), 1–31.  
21  
22  
23 335 Environment Agency. (2016a). Flood risk management plans (FRMPs): 2015 to 2021. Retrieved  
24 336 November 6, 2017, from [https://www.gov.uk/government/collections/flood-risk-management-](https://www.gov.uk/government/collections/flood-risk-management-plans-frmps-2015-to-2021)  
25 [plans-frmps-2015-to-2021](https://www.gov.uk/government/collections/flood-risk-management-plans-frmps-2015-to-2021)  
26 337  
27  
28 338 Environment Agency. (2016b). How to model and map catchment processes when flood risk  
29 339 management planning, 77. Retrieved from  
30 340 [https://www.gov.uk/government/publications/how-to-model-and-map-catchment-processes-](https://www.gov.uk/government/publications/how-to-model-and-map-catchment-processes-when-flood-risk-management-planning)  
31 [when-flood-risk-management-planning](https://www.gov.uk/government/publications/how-to-model-and-map-catchment-processes-when-flood-risk-management-planning)  
32 341  
33  
34  
35 342 Environment Food and Rural Affairs Committee. (2016). *Future flood prevention Second Report of*  
36 343 *Session 2016–17*. Retrieved from  
37 344 <https://publications.parliament.uk/pa/cm201617/cmselect/cmenvfru/115/115.pdf>  
38  
39  
40  
41 345 Evans et al. (2004). Foresight Future Flooding. *Office of Science and Technology*, 1–59. Retrieved  
42 346 from [http://www.bis.gov.uk/foresight/our-work/projects/published-projects/flood-and-](http://www.bis.gov.uk/foresight/our-work/projects/published-projects/flood-and-coastal-defence)  
43 [coastal-defence](http://www.bis.gov.uk/foresight/our-work/projects/published-projects/flood-and-coastal-defence)  
44 347  
45  
46  
47 348 Forbes, H., Ball, K., & McLay, F. (2015). *Natural Flood Management Handbook*.  
48 349 <https://doi.org/10.13140/RG.2.1.4956.1444>  
49  
50  
51 350 Hooke, J. M. (2015). Variations in flood magnitude-effect relations and the implications for flood risk  
52 351 assessment and river management. *Geomorphology*, 251, 91–107.  
53 352 <https://doi.org/10.1016/j.geomorph.2015.05.014>  
54  
55  
56  
57 353 Howarth, W. (2017). Integrated Water Resources Management and reform of flood risk  
58 354 management in England. *Journal of Environmental Law*, 29(2), 355–365.  
59 355 <https://doi.org/10.1093/jel/eqx015>  
60

- 1  
2  
3 356 Howe, J., & White, I. (2003). Flooding, pollution and agriculture. *International Journal of*  
4 *Environmental Studies*, 60(1), 19–27. <https://doi.org/10.1080/00207230304746>  
5 357  
6  
7 358 Hurlimann, A., & Wilson, E. (2018). Sustainable urban water management under a changing climate:  
8 The role of spatial planning. *Water (Switzerland)*, 1–22. <https://doi.org/10.3390/w10050546>  
9 359  
10  
11 360 Jones, P., & Macdonald, N. (2007). Getting it wrong first time: Building an interdisciplinary research  
12 relationship. *Area*, 39(4), 490–498. <https://doi.org/10.1111/j.1475-4762.2007.00767.x>  
13 361  
14  
15 362 Jongman, B., Hochrainer-Stigler, S., Feyen, L., Aerts, J. C. J. H., Mechler, R., Botzen, W. J. W., ... Ward,  
16 P. J. (2014). Increasing stress on disaster-risk finance due to large floods. *Nature Climate*  
17 363 *Change*, 4(4), 264–268. <https://doi.org/10.1038/nclimate2124>  
18 364  
19  
20  
21 365 Kidd, S., & Shaw, D. (2007). Integrated Water Resource Management and Institutional Integration:  
22 Realising the Potential of Spatial Planning in England. *The Geographical Journal*, Vol. 173, No.4,  
23 366 *December 2007*, Pp. 312-329, 173(4), 312–329.  
24 367  
25  
26 368 Kundzewicz, Z. W., Kanae, S., Seneviratne, S. I., Handmer, J., Nicholls, N., Peduzzi, P., ... Sherstyukov,  
27 B. (2014). Flood risk and climate change: global and regional perspectives. *Hydrological*  
28 369 *Sciences Journal*, 59(1), 1–28. <https://doi.org/10.1080/02626667.2013.857411>  
29 370  
30  
31  
32 371 Leach, W. D., & Pelkey, N. W. (2001). Making Watershed Partnerships Work : a Review Empirical  
33 372 Literature. *Journal Of Water Resources Planning And Management*, 127(DECEMBER), 378–385.  
34  
35  
36 373 Lewis, B. (2014). House of Commons : Written Statement ( HCWS50 ), 2014–2016.  
37  
38  
39 374 Liao, K.-H. (2012). A Theory on Urban Resilience to Floods - A Basis for Alternative Planning Practices.  
40 375 *Ecology and Society*, 17(4), 15. <https://doi.org/10.5751/ES-05231-170448>  
41  
42  
43 376 Lieberherr, E., & Truffer, B. (2015). The impact of privatization on sustainability transitions: A  
44 377 comparative analysis of dynamic capabilities in three water utilities. *Environmental Innovation*  
45 378 *and Societal Transitions*, 15, 101–122. <https://doi.org/10.1016/j.eist.2013.12.002>  
46  
47  
48 379 Maas, J. (2006). Green space, urbanity, and health: how strong is the relation? *Journal of*  
49 380 *Epidemiology & Community Health*, 60(7), 587–592. <https://doi.org/10.1136/jech.2005.043125>  
50  
51  
52 381 McIntyre, N., & Thorne, C. (2013). *Land use management effects on flood flows and sediments -*  
53 382 *guidance on prediction*.  
54  
55  
56 383 Medcalf, K., Small, N., Finch, C., Williams, J., Blair, T., Haines-Young, R., Potschin, M. & Parker, J.  
57 384 (2014). Further development of a spatial framework for mapping ecosystem services  
58 385 development of a spatial framework for mapping ecosystem services, (514). Retrieved from



1  
2  
3 386 <http://jncc.defra.gov.uk>  
4

5 387 Metcalfe, P., Beven, K., Hankin, B., & Lamb, R. (2017). A modelling framework for evaluation of the  
6  
7 388 hydrological impacts of nature-based approaches to flood risk management, with application  
8  
9 389 to in-channel interventions across a 29-km<sup>2</sup> scale catchment in the United Kingdom.

10 390 *Hydrological Processes*, 31(9), 1734–1748. <https://doi.org/10.1002/hyp.11140>  
11

12  
13 391 Miller, J. D., Kim, H., Kjeldsen, T. R., Packman, J., Grebby, S., & Dearden, R. (2014). Assessing the  
14  
15 392 impact of urbanization on storm runoff in a peri-urban catchment using historical change in  
16  
17 393 impervious cover. *Journal of Hydrology*, 515, 59–70.

18 394 <https://doi.org/10.1016/j.jhydrol.2014.04.011>  
19

20 395 National Trust. (2015). From source to sea - Natural Flood Management - The Holnicote Experience,  
21  
22 396 (March), 28.

23  
24 397 Neuhold, C. (2014). *Links between the Floods Directive and the Water Framework Directive. PLANALP*  
25  
26 398 *Conference*. <https://doi.org/10.2779/71412>  
27

28 399 Newig, J., Challies, E., Jager, N., & Kochskämper, E. (2014). What role for public participation in  
29  
30 400 implementing the EU floods directive? A comparison with the water framework directive, early  
31  
32 401 evidence from Germany and a research agenda. *Environmental Policy and Governance*, 24(4),  
33  
34 402 275–288. <https://doi.org/10.1002/eet.1650>

35  
36 403 Nisbet, T., Marrington, S., Thomas, H., Broadmeadow, S., & Valantin, G. (2011). Defra FCERM Multi-  
37  
38 404 objective Flood Management Demonstration project. *Forestry Research*, (April), 1–6.

39  
40 405 Pattison, I., & Lane, S. N. (2012). The link between land-use management and fluvial flood risk.  
41  
42 406 *Progress in Physical Geography*, 36(1), 72–92. <https://doi.org/10.1177/0309133311425398>  
43

44 407 Pitt, M. (2008). The Pitt Review: Learning Lessons from the 2007 Floods. *Floods Review*, 19,.  
45  
46 408 <https://doi.org/10.1007/s13398-014-0173-7.2>

47  
48 409 Postnote. (2016). *Green Space and Health*. Retrieved from file:///C:/Users/tajw/Downloads/POST-  
49  
50 410 PN-0538.pdf

51  
52 411 Rahaman, M. M., Varis, O., & Kajander, T. (2004). EU Water Framework Directive vs. Integrated  
53  
54 412 Water Resources Management: The seven mismatches. *International Journal of Water*  
55  
56 413 *Resources Development*, 20(4), 565–575. <https://doi.org/10.1080/07900620412331319199>

57  
58 414 Ran, J., & Nedovic-Budic, Z. (2016). Integrating spatial planning and flood risk management: A new  
59  
60 415 conceptual framework for the spatially integrated policy infrastructure. *Computers*,

- 1  
2  
3 416 *Environment and Urban Systems*, 57, 68–79.  
4  
5 417 <https://doi.org/10.1016/j.compenvurbsys.2016.01.008>  
6  
7 418 Redfern, T. W., Macdonald, N., Kjeldsen, T. R., Miller, J. D., & Reynard, N. (2016). Current  
8  
9 419 understanding of hydrological processes on common urban surfaces. *Progress in Physical*  
10  
11 420 *Geography*, 40(5), 699–713. <https://doi.org/10.1177/0309133316652819>  
12  
13 421 Richter, S., Völker, J., Borchardt, D., & Mohaupt, V. (2013). The Water Framework Directive as an  
14  
15 422 approach for Integrated Water Resources Management: Results from the experiences in  
16  
17 423 Germany on implementation, and future perspectives. *Environmental Earth Sciences*, 69(2),  
18  
19 424 719–728. <https://doi.org/10.1007/s12665-013-2399-7>  
20  
21 425 Riley, M., Sangster, H., Smith, H., Chiverrell, R., & Boyle, J. (2018). Will farmers work together for  
22  
23 426 conservation? The potential limits of farmers' cooperation in agri-environment measures. *Land*  
24  
25 427 *Use Policy*, 70(February 2018), 635–646. <https://doi.org/10.1016/j.landusepol.2017.10.049>  
26  
27 428 Robins, L., Burt, T. P., Bracken, L. J., Boardman, J., & Thompson, D. B. A. (2017). Making water policy  
28  
29 429 work in the United Kingdom: A case study of practical approaches to strengthening complex,  
30  
31 430 multi-tiered systems of water governance. *Environmental Science and Policy*, 71, 41–55.  
32  
33 431 <https://doi.org/10.1016/j.envsci.2017.01.008>  
34  
35 432 Rouillard, J. J., Heal, K. V., Ball, T., & Reeves, A. D. (2013). Policy integration for adaptive water  
36  
37 433 governance: Learning from Scotland's experience. *Environmental Science and Policy*, 33, 378–  
38  
39 434 387. <https://doi.org/10.1016/j.envsci.2013.07.003>  
40  
41 435 Smith, Laurence, Porter, Keith, Hiscock, Kevin., P. M. J. (2015). *Catchment and River Basin*  
42  
43 436 *Management Integrating Science and Governance*. Routledge.  
44  
45 437 Spray, C., Ball, T., & Rouillard, J. (2009). Bridging the water law, policy, science interface: flood risk  
46  
47 438 management in Scotland. *Water Law*, 20(January), 165–174.  
48  
49 439 Stevens, R., & Ogunyoye, F. (2012). Costs and Benefits of Sustainable Drainage Systems, (July), 1–24.  
50  
51 440 Retrieved from [https://www.theccc.org.uk/archive/aws/ASC/2012 report/Royal Haskoning](https://www.theccc.org.uk/archive/aws/ASC/2012%20report/Royal%20Haskoning%20Costs%20and%20Benefit%20of%20SuDS%20Final%20Report.pdf)  
52  
53 441 [Costs and Benefit of SuDS Final Report.pdf](https://www.theccc.org.uk/archive/aws/ASC/2012%20report/Royal%20Haskoning%20Costs%20and%20Benefit%20of%20SuDS%20Final%20Report.pdf)  
54  
55 442 Thorne, C. (2014). Geographies of UK flooding in 2013/4. *Geographical Journal*, 180(4), 297–309.  
56  
57 443 <https://doi.org/10.1111/geoj.12122>  
58  
59 444 Walker, L., Wright, G., Digman, C., Glerum, J., & Ashley, R. (2011). Are UK professionals predisposed  
60  
445 to retrofit more sustainable surface water management measures ? *12th International*  
446 *Conference on Urban Drainage*, (September), 11–16.

- 1  
2  
3 447 Water UK. (2015). Sustainable drainage - refining the approach. Retrieved November 17, 2017, from  
4  
5 448 <https://www.water.org.uk/news-water-uk/latest-news/sustainable-drainage-refining-approach>  
6  
7 449 Watson, N. (2015). Factors Influencing the Frames and Approaches of Host Organizations for  
8  
9 450 Collaborative Catchment Management in England. *Society and Natural Resources*, 28(4), 360–  
10 451 376. <https://doi.org/10.1080/08941920.2014.945059>  
12  
13 452 Waylen, K. A., Holstead, K. L., Colley, K., & Hopkins, J. (2017). Challenges to enabling and  
14 453 implementing Natural Flood Management in Scotland. *Journal of Flood Risk Management*, 1–  
15 454 12. <https://doi.org/10.1111/jfr3.12301>  
17  
18 455 Werritty, A. (2006). Sustainable flood management: oxymoron or new paradigm? *Area*, 38(1), 16–23.  
19 456 <https://doi.org/10.1111/j.1475-4762.2006.00658.x>  
21  
22 457 Westcountry Rivers Trust. (n.d.). Participatory Ecosystem Services Visualisation Framework - Making  
23 458 effective use of data & evidence to inform catchment management planning, 1–42.  
25  
26 459 Woodward, M., Gouldby, B., Kapelan, Z., Khu, S., Townend, I., & Park, H. (2009). the Use of Real  
27 460 Options in Optimum Flood Risk, (2008), 1–6.  
29  
30 461

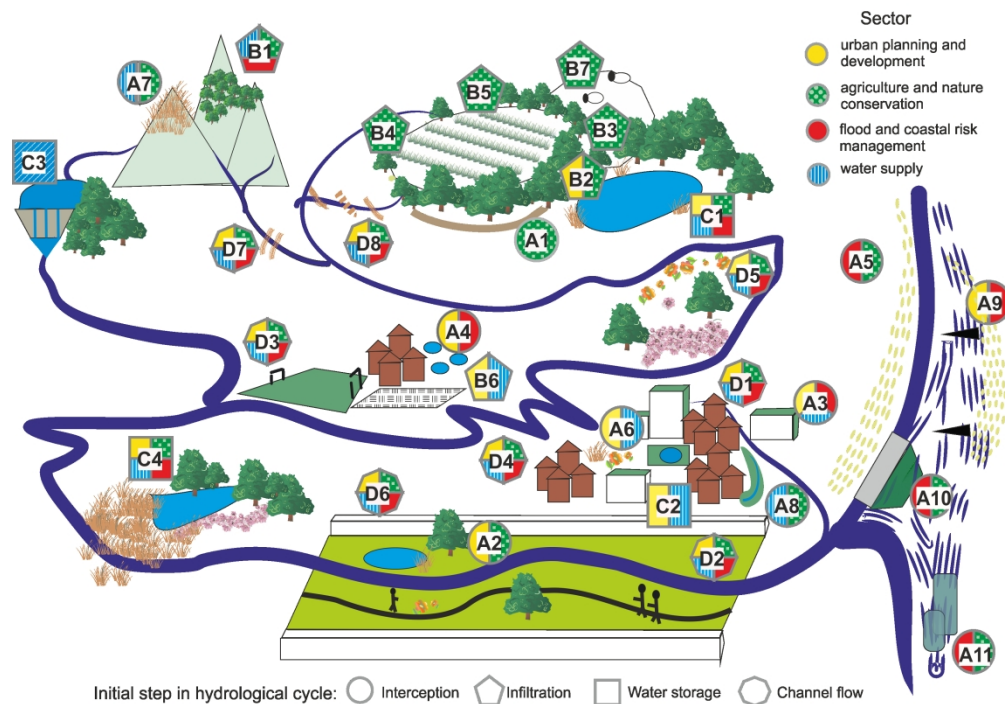


Figure 1. Catchment wide NFM interventions categorised as the initial step in the hydrological cycle – Interception: A1 bunded ditches, A2 vegetative cover, A3 green roofs and walls, A4 interception ponds, A5 managed realignment, A6 rain gardens, A7 restoring peatlands, A8 swales, A9 beach nourishment, A10 habitat promotion, A11 reef creation Infiltration: B1 woodlands, B2 filter/buffer strips, B3 hedgerows, B4 managing soil quality, B5 no and low till agriculture, B6 permeable paving, B7 reduced stocking density Water storage: C1 ponds, C2 rainwater harvesting, C3 reservoirs, C4 wetlands and reed beds Channel flow: D1 de-culverting, D2 increase channel roughness, D3 regulated washlands, D4 re-meandering, D5 restore functioning floodplain, D6 setting back flood defences, D7 woody material dams D8 species reintroduction (e.g. beavers). Each intervention uses a number of hydrological processes to slow the flow of water for example interception, infiltration and water storage in wetlands and surrounding vegetative cover will result in reduced surface run-off.