

1 **GLOWORM-FL: a simulation model of the effects of climate and climate**  
2 **change on the free-living stages of gastro-intestinal nematode parasites of**  
3 **ruminants**

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13

14 **Abstract**

15 Gastrointestinal nematodes are important parasites of livestock and wildlife  
16 worldwide, causing mortality and morbidity, regulating host populations and  
17 threatening food security through reduced productivity of ruminant livestock. A  
18 significant part of the life-cycle of most GINs is completed outside of the host. GINs  
19 are therefore susceptible to changes in climate, and evidence of climate-driven  
20 changes in the phenology of GINs and the seasonal incidence of disease already  
21 exists. A modelling framework, GLOWORM-FL was developed to predict changes in  
22 the seasonal dynamics of the free-living stages of trichostrongylid GINs on pasture  
23 as a first step towards evaluating potential mitigation strategies. The general model  
24 framework was parameterised and validated for three GIN species that infect a  
25 range of ruminants worldwide: *Haemonchus contortus*, *Teladorsagia circumcincta*  
26 and *Ostertagia ostertagi*. The model builds significantly on previous models of GIN  
27 population dynamics by incorporating the behaviour of nematodes in response to  
28 climate variability, facilitated by recent advances in our understanding of the ecology  
29 of GINs. Simulations using historical and predicted future climatic data for a  
30 temperate region reveal the potential for an increase in annual infection pressure of  
31 *H. contortus* and *T. circumcincta* in small ruminants as increasing temperatures  
32 accelerate development and remove constraints on the development of *H. contortus*  
33 during the winter months. In contrast, a significant decrease in annual infection  
34 pressure is predicted for *O. ostertagi* in cattle due to accelerated development being  
35 offset by rapid mortality at higher temperatures. A similar trade-off is predicted during  
36 the summer months for *H. contortus* and *T. circumcincta* resulting in complex  
37 seasonal dynamics of the availability of infective stages on pasture. These changes  
38 could have significant impacts on the seasonal incidence and pathology of infection

39 by GINs. GLOWORM-FL therefore provides an important tool to predict the seasonal  
40 risk of transmission of GINs and will aid in the design of climate-driven, risk-based  
41 GIN control strategies.

42 **Keywords:** gastrointestinal nematode; climate change; population dynamics;  
43 nematode behaviour; ruminant; parasite ecology

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## 46 **1. Introduction**

47 Trichostrongyloid gastrointestinal nematodes (GINs) are a major cause of mortality  
48 and morbidity in livestock (e.g. Allonby and Urquhart, 1975), threatening food  
49 security via constraints on productivity (Fitzpatrick, 2013). Costs of GINs have been  
50 estimated at 84 million pounds sterling (105 million Euros) annually for the sheep  
51 industry in the UK alone (Nieuwhof and Bishop, 2007), although effects of infection  
52 on farm economics can be complex and difficult to estimate (van der Voort et al.,  
53 2013).

54 Adult trichostrongylid GINs inhabit the gastrointestinal system of a range of host  
55 species including ruminants (Allonby and Urquhart, 1975; Morgan et al., 2005),  
56 lagomorphs (Newey et al., 2005) and birds (Hudson et al., 1998). Eggs are  
57 deposited in the environment in faeces, where they develop to infective larvae, which  
58 then move onto herbage. Larvae are ingested by the host during grazing and  
59 complete their life-cycle in the host (Anderson, 2000). The development, survival and  
60 behaviour of the free-living stages and thus the availability of infective stages for  
61 transmission is highly dependent on weather and micro-climatic conditions (Khadijah  
62 et al., 2013a; Morgan and van Dijk, 2012; O'Connor et al., 2008, 2007, 2006;  
63 Reynecke et al., 2011; Rose, 1963, 1961; van Dijk and Morgan, 2008, 2011). There  
64 is evidence that recent increases in temperature in the UK have resulted in changes  
65 in the phenology of GINs on pasture (Gethings et al. submitted) and in the incidence  
66 of disease due to GIN infection (parasitic gastroenteritis) in livestock (van Dijk et al.,  
67 2008). As a result, the potential impact of climate change on GIN-host dynamics is of  
68 increasing concern (Rose et al., 2014; van Dijk and Morgan, 2010; van Dijk et al.,  
69 2010).

70 Predicting climate-driven changes in the seasonal availability of free-living GIN  
71 infective stages is the first step to evaluating the potential impact of climate change  
72 on GIN infections in livestock and wildlife and developing sustainable strategies to  
73 control GINs and mitigate any increased transmission risk. These baseline  
74 predictions of infection pressure can then be integrated with patterns of host  
75 availability to evaluate the seasonal risk of transmission (e.g. Morgan et al., 2006).  
76 However, predicting the response of GINs to climate change is complicated by  
77 nonlinear relationships and interactions between climate, development and survival  
78 (Molnár et al., 2013), and the system necessitates parsimonious predictive models  
79 that balance sufficient biological detail with experimentally verifiable parameters  
80 (Morgan, 2013) and computational efficiency.

81 Numerous gastrointestinal nematode models have been developed over previous  
82 decades (reviewed elsewhere by Cornell, 2005; Roberts, 1995; Smith and Grenfell,  
83 1994). Many are deliberately simple in order to explore model behaviour and system  
84 dynamics (Cornell et al., 2004; Grenfell, 1992; Louie et al., 2005; Roberts and  
85 Heesterbeek, 1995). Others include more biological detail in order to address  
86 specific questions (Grenfell et al., 1987; Laurenson et al., 2011; Learmount et al.,  
87 2006; Leathwick et al., 1995, 1992; Smith et al., 1987). However, climate-dependent  
88 life-history parameters that determine the availability of infective stages on pasture  
89 are often set at a constant rate (Laurenson et al., 2013). Furthermore, although  
90 many models incorporate climate-dependence (Grenfell et al., 1987; Molnár et al.,  
91 2013) and stage-specific mortality and development rates, to the authors' knowledge  
92 no model explicitly incorporates movement of infective larvae between soil and  
93 herbage nor addresses moisture-limitations on migration between faeces and  
94 pasture (herbage and soil combined). Detail such as this will become increasingly

95 important as increases in the frequency of extreme events such as drought and  
96 heavy rainfall are predicted into the late 21<sup>st</sup> century (IPCC, 2013).

97 The model framework presented here, GLOWORM-FL, builds on the work of  
98 Grenfell et al. (1987) and Smith (1990) by incorporating recent advances in our  
99 understanding of the behaviour and ecology of GINs on pasture to predict the  
100 climate-dependent seasonal dynamics of GIN infection pressure. The model  
101 provides a generic framework that can be applied to a range of GIN species. To  
102 demonstrate the flexibility of the framework and methods for data-driven parameter  
103 estimation, the model is parameterised and validated for three trichostrongylid GIN  
104 species - *Haemonchus contortus*, *Teladorsagia circumcincta* and *Ostertagia*  
105 *ostertagi* - and used to simulate the seasonal dynamics of the availability of infective  
106 stages on pasture under scenarios of likely climate change, independent of host  
107 factors.

108 The three species of GIN chosen here are of economic importance to the ruminant  
109 livestock industry worldwide, but also have a broad host range and infect free-  
110 ranging ruminants. The haematophagous abomasal nematode, *H. contortus* is highly  
111 pathogenic. Chronic infections in sheep may result in anaemia and death (Allonby  
112 and Urquhart, 1975). The abomasal nematodes *T. circumcincta* and *O. ostertagi* are  
113 responsible for significant production losses in the ruminant livestock industry  
114 (Charlier et al., 2009; Nieuwhof and Bishop, 2007). Anthelmintic resistance is  
115 increasingly widespread in all three species in livestock (De Graef et al., 2013;  
116 Kaplan and Vidyashankar, 2012; Papadopoulos et al., 2012; Sutherland and  
117 Leathwick, 2011) and has been recorded in *H. contortus* in wild deer (Chintoan-Uta  
118 et al., 2014). A better understanding of the population ecology of these parasites is

119 therefore needed to underpin the development of alternative control strategies  
120 against the backdrop of climate change and anthelmintic resistance.

121

## 122 2. Materials and Methods

### 123 2.1 GLOWORM-FL model framework

124 The model framework is based on the general life-cycle of the free-living stages of  
125 trichostrongylid gastrointestinal nematodes (Figure 1). Eggs ( $E$ ) develop to third  
126 stage larvae in the faeces ( $L3_f$ ) via the pre-infective larval stages ( $L$ ), and are subject  
127 to stage-specific mortality rates ( $\mu_i$ ). As pre-infective larval stages (first stage larvae,  
128 L1, and second-stage larvae, L2) are not separated, the model can be applied to  
129 trichostrongylids that hatch as first stage larvae (e.g. *Haemonchus* spp.,  
130 *Teladorsagia* spp. and *Ostertagia* spp.; Anderson, 2000), or second stage larvae  
131 (*Marshallagia marshalli*; Carlsson et al., 2013).

$$\frac{dE}{dt} = -(\mu_1 + 2\delta)E + E_{new}C \quad (1)$$

$$\frac{dL}{dt} = -(\mu_2 + 2\delta)L + 2\delta E \quad (2)$$

$$\frac{dL3_f}{dt} = -(\mu_3 + m_1)L3_f + 2\delta L \quad (3)$$

132 As development from egg to L3 in faeces is divided over two stages, the development  
133 rate ( $\delta$ ) is doubled. The framework tracks numbers of overlapping cohorts of  
134 nematodes, and so new eggs deposited on pasture ( $E_{new}$ ) join the pool of existing  
135 eggs.

136 Previous models of GIN free-living stages either model total L3 on pasture and  
137 implicitly assume that, once developed, L3 are available for transmission (Grenfell et  
138 al., 1987; Learmount et al., 2006), or separate L3 in faeces from L3 on pasture and  
139 apply a constant horizontal migration rate (Grenfell et al., 1986; Smith, 1990). Here,



140 L3 in faeces actively migrate from faeces onto pasture at a climate-dependent  
141 horizontal migration rate ( $m_1$ ).

142 Once on pasture, L3 can be recovered from both soil/mat layer and herbage. Although  
143 Grenfell et al. (1986) included losses of L3 in soil due to soil moisture deficit or wash-  
144 down during rainfall in their model, movement between soil and herbage was not  
145 considered. Experiments have demonstrated the potential for bi-directional movement  
146 of trichostrongylid L3 between soil and herbage (Krecek and Murrell, 1988; Rose and  
147 Small, 1985) and that there is random movement between the soil and herbage (van  
148 Dijk and Morgan 2011). Therefore, L3 on pasture ( $L3_p$ ) are assumed to reside in either  
149 the soil and vegetation mat layer ( $L3_s$ ) or on herbage ( $L3_h$ ). In order to simulate  
150 random, bi-directional movement between herbage and the soil reservoir, substrate-  
151 specific mortality rates ( $\mu_4$ ,  $\mu_5$ ) are applied to the proportion of larvae estimated to  
152 reside in soil and on herbage respectively, dependant on a vertical migration  
153 parameter ( $m_2$ ).

$$\frac{dL3_p}{dt} = -\mu_4 \left( (1 - m_2) L3_p \right) - \mu_5 (m_2 L3_p) + m_1 L3_f \quad (4)$$

154 State variables and parameter definitions are listed in Table 1. The model was  
155 implemented in R (R Core Team, 2013) using the *Isoda* function in the *deSolve*  
156 package (Soetaert et al., 2010). *Isoda* uses an Adams-BDF (backward differentiation  
157 formulae) adaptive integration method that detects the stiffness of the problem  
158 throughout the simulation and switches between Adams and BDF integration  
159 accordingly (Soetaert et al., 2010). The model returns daily output but the time steps  
160 used for integration are not known prior to simulation when using the Adams-BDF  
161 integration method. Therefore, time-series of variable climate-dependent rates e.g.  
162 temperature-dependent development rates, were generated prior to simulation and

163 introduced by interpolation using the *approxfun* function (Soetaert et al., 2012). New  
164 eggs were deposited using the “events” argument of the *Isoda* function (Soetaert et  
165 al., 2012). Model output is numbers of individuals per unit area e.g. per hectare, and  
166 is therefore independent of herbage density.

## 167 **2.2 Parameter estimates**

168 The model was parameterised for three trichostrongylid GINs that infect ruminants:  
169 *Haemonchus contortus*, *Teladorsagia circumcincta* and *Ostertagia ostertagi* (Figures  
170 2-4; Table 2).

### 171 **2.2.1 Temperature dependent development and mortality**

172 Temperature-dependent instantaneous daily rates were estimated for development  
173 from egg to L3 in faeces and stage- and substrate-specific mortality using data from  
174 experiments that reported the proportions of individuals developed (for development  
175 rates) or surviving (for mortality rates) at discrete intervals and at a range of constant  
176 temperatures (Table 2).

177 Instantaneous daily rates were first estimated for each constant temperature from  
178 the reported time to 50% development (D50) or time to 50% mortality (M50)  
179 as  $-\ln(0.5)/D50$  or  $-\ln(0.5)/M50$ . If these data were unavailable, rates were estimated  
180 in one of three ways: 1) using the proportion remaining at a single sampling interval,  
181 as  $-\ln(\text{proportion remaining})/\text{days}$ ; 2) using the mean of the minimum and maximum  
182 development or mortality times, or; 3) by linear regression of the transformed  
183 proportions of individuals developed or surviving over time as described by Azam et  
184 al. (2012). Where 100% mortality was observed within 24 hours, an instantaneous  
185 mortality rate of 1 was applied.

186 Linear models were then fitted to the instantaneous daily rates at a range of  
187 temperatures, yielding a regression equation that could be used to estimate daily  
188 rates dependent on time-series of observed temperature data (Table 2). For  
189 development rates, which increase linearly as a function of temperature, simple  
190 linear regression was used. Mortality rates are highest at extreme high and low  
191 temperatures, therefore polynomial models were fitted to the log-transformed  
192 instantaneous mortality rates. Rates were limited between 0 and 1 where necessary.

193 Data were only used to estimate the mortality rates of pre-infective stages if the  
194 temperatures were low enough to preclude development, or high enough that  
195 mortality occurred prior to development to the next larval stage (e.g. Todd et al.,  
196 1976). Development of L3 is arrested until they are ingested by the host. Therefore a  
197 range of temperatures could be used to estimate the substrate-specific mortality  
198 rates of L3.

199 The mortality rate of L3 in soil ( $\mu_4$ ) for all GIN species was estimated using  
200 observations on the mortality of L3 in water, which for *H. contortus* and *T.*  
201 *circumcincta*, provided point estimates of instantaneous daily mortality rates similar  
202 to those reported by van Dijk and Morgan (2011) in soil at 20-24°C (Figures 2-3).

203 Exposure to UV irradiation increases the mortality of trichostrongyloid L3 in water  
204 (van Dijk et al., 2009) and the estimates of mortality in soil (Table 2; van Dijk and  
205 Morgan, 2011) are considerably lower than estimates of mortality on pasture  
206 (Grenfell et al., 1986). Therefore it is assumed that the mortality rate of L3 on  
207 herbage is higher than in soil, and the mortality of L3 in faeces ( $\mu_3$ ) was used as a  
208 proxy for L3 mortality on herbage ( $\mu_5$ ; Table 2).

209 No data were available to estimate the mortality of pre-infective larvae ( $\mu_2$ ) and L3 in  
210 faeces ( $\mu_3$ ) for *T. circumcincta* and *O. ostertagi*, Therefore the same mortality rate  
211 was used for eggs ( $\mu_1$ ) and pre-infective larvae ( $\mu_2$ ; Table 2). To estimate L3  
212 mortality in faeces the temperature-dependent mortality of *O. ostertagi* in water  
213 (used to estimate  $\mu_4$ ) was compared with point estimates of mortality of *O. ostertagi*  
214 and *Cooperia oncophora* in cow manure (Persson, 1974a). The instantaneous daily  
215 mortality rates at 20°C and 3°C were estimated using Persson's data. As there was  
216 significant variability between sampling intervals the instantaneous mortality rate was  
217 calculated for each sampling interval and the mean estimated from these rates.  
218 Based on these analyses L3 mortality in soil is 4.9-18.5 times lower than in faeces.  
219 Therefore, in the absence of data to directly estimate the mortality rate of *T.*  
220 *circumcincta* and *O. ostertagi* L3 in faeces, it is estimated that  $\mu_3 = 10\mu_4$ , within the  
221 limits of 0 and 1 (Figures 3-4).

## 222 **2.2.2 Moisture limitations and differences between host species**

223 Moisture limitations on the availability of GIN infective stages are primarily mediated  
224 through changes in faecal moisture content (FMC; Mauleon and Gruner, 1984;  
225 Rossanigo and Gruner, 1995). There are significant differences in faeces structure  
226 and drying rates between host species. Sheep faecal pellets tend to dry rapidly  
227 following deposition, whereas the decrease in cow pat FMC is more gradual  
228 (Mauleon and Gruner, 1984). It is therefore necessary to not only parameterise the  
229 model for different nematode species, but also different host species. Here, we  
230 consider moisture limitations on *H. contortus* and *T. circumcincta* infecting sheep or  
231 other ruminants with a similar faecal pellet structure, and *O. ostertagi* infecting cattle.

### 232 **2.2.2.1 Moisture limitations on development success**

233 In addition to temperature limitations, development success (the proportion of eggs  
234 that develop to L3) is also a function of faecal moisture content (Rossanigo and  
235 Gruner, 1995). To impose faecal moisture limitations on development and mortality  
236 of *H. contortus* and *T. circumcincta* without explicitly modelling FMC, cumulative  
237 precipitation divided by cumulative potential evapotranspiration (referred to as  
238 cumulative P/E) is estimated for a species-specific critical period following deposition  
239 of eggs after O'Connor et al. (2008), who observed a strong positive relationship  
240 between cumulative P/E and the FMC of sheep faecal pellets. If cumulative P/E < 1  
241 then the number of new eggs ( $E_{new}$ ) entering the pool of eggs in faeces is reduced by  
242 an amount specified by the correction factor parameter,  $C$ .

243 O'Connor et al. (2008) observed a significant decrease in the development success  
244 of *H. contortus* where cumulative P/E fell below 1 within 4 days of deposition of eggs.  
245 Khadijah et al. (2013a) recovered maximum *H. contortus* L3 from faecal pellets and  
246 soil when simulated rainfall was applied between -1 and 2 days post deposition of  
247 faeces containing eggs and concluded that faecal moisture 48-72 hours post  
248 deposition was important for development success. No L3 were recovered from un-  
249 watered controls. Similar data were not available for *T. circumcincta*. However,  
250 Khadijah et al. (2013a) note that for *Trichostrongylus colubriformis*, faecal moisture  
251 in the period 72-96h post deposition is important for development success. This  
252 period is likely to be extended for *T. circumcincta* which is more resistant to  
253 desiccation than *T. colubriformis*. A lower faecal moisture content (FMC) threshold  
254 was observed for *T. circumcincta* development (yielding  $\geq 1$  L3 per 100 eggs) than *T.*  
255 *colubriformis* (25% and 35% respectively; Rossanigo and Gruner, 1995). The critical  
256 periods for *H. contortus* and *T. circumcincta* were therefore identified as 4 days and  
257 7 days post deposition of eggs respectively (Khadijah et al., 2013b; O'Connor et al.,

258 2008). The cumulative P/E for development success is referred to as  $P/E_4$  for *H.*  
259 *contortus* and  $P/E_7$  for *T. circumcincta*.

260 A protective surface crust forms on cow pats soon after deposition. It is therefore  
261 assumed that moisture is not limiting for GIN development within cattle faeces at  
262 lower FMCs (Rose, 1961).

### 263 **2.2.2.2 Moisture limitations on the translation of L3 onto pasture**

264 Laboratory observations on the migration of *H. contortus* and *T. colubriformis* L3  
265 indicate that, similar to development success, moisture limitations on horizontal  
266 migration are mediated by faecal moisture content, which varies as a result of  
267 interacting microclimatic factors (Khadijah et al., 2013a, 2013b; O'Connor et al.,  
268 2008, 2007; van Dijk and Morgan, 2011). Few data were available to estimate the  
269 temperature- and moisture-dependent horizontal migration rate of infective larvae  
270 from faeces onto pasture. Therefore, data in the published literature were  
271 supplemented with laboratory experiments to derive heuristic estimates for horizontal  
272 migration under: 1) optimal moisture conditions (sufficient rainfall); 2) sub-optimal  
273 moisture conditions (insufficient rainfall but sufficient FMC) and; 3) moisture-limiting  
274 conditions (low FMC and insufficient rainfall).

275 Horizontal migration of GINs has been observed from cow pats following 1.6mm of  
276 simulated rainfall (Grønvold and Høgh-Schmidt, 1989) and from sheep faecal pellets  
277 following 2mm of simulated rainfall (this study). Furthermore, horizontal migration  
278 rates of *H. contortus* were not significantly influenced by the amount of rainfall  
279 between 4mm and 8mm (Wang et al., 2014). Therefore, optimal moisture was  
280 defined as days where total precipitation  $\geq 2$ mm.

281 A daily horizontal migration rate of 0.06 (S.D. 0.057) applied for *O. ostertagi* based  
282 on the number of L3 recovered by Grønvold and Høgh-Schmidt (1989) from within  
283 and outside of cow pats after 1.6-1.7mm simulated rainfall was applied to pats with  
284 FMCs of 54-66%.

285 To estimate the daily horizontal migration rate for *H. contortus* and *T. circumcincta*  
286 faeces containing either *H. contortus* or *T. circumcincta* eggs provided by Moredun  
287 Research Institute, Edinburgh, UK were incubated at 20°C for 7 days and then  
288 allowed to dry at room temperature for varying amounts of time to obtain pellets with  
289 variable initial faecal moisture content (FMC). Three replicates of 3g (~6 pellets)  
290 were subjected to approximately 2mm simulated rainfall and after 24 hours L3 that  
291 had migrated out of faeces and L3 remaining in the faeces were recovered and  
292 enumerated (Wang et al., 2014). The recovery efficiency of extra-pellet and intra-  
293 pellet L3 were determined to be 84% (S.D. 3%) and 74% (S.D. 7%) respectively by  
294 placing a known number of L3 in the cup used to contain L3 that had migrated out of  
295 faeces (Wang et al., 2014) and seeding faeces with a known number of L3. The  
296 weight of each pellet was recorded before and immediately after the rainfall event,  
297 and after drying in an oven, to estimate the FMC at each stage.

298 The proportion of L3 that had migrated out of faeces was calculated as:  
299  $\text{extra-pellet L3}/(\text{extra-pellet L3} + \text{intra-pellet L3})$ , corrected for recovery efficiency.  
300 Mean daily horizontal migration rates of 0.25 (S.D. 0.11) and 0.21 (S.D. 0.44) were  
301 observed for *H. contortus* and *T. circumcincta*. FMCs prior to rainfall were 3-61% (*H.*  
302 *contortus*) and 7-34% (*T. circumcincta*), increasing to 45-73% (*H. contortus*) and 39-  
303 56% (*T. circumcincta*) after simulated rainfall.

304 *O. ostertagi* L3 were only observed on the surface of experimental pats that had  
305 been watered (Grønvold and Høgh-Schmidt, 1989) suggesting that rainfall is  
306 required to moisten the protective surface crust sufficiently to allow migration.  
307 Therefore, estimates of horizontal migration of *O. ostertagi* under sub-optimal  
308 moisture conditions were not considered. In contrast, small numbers of extra-pellet  
309 *H. contortus* and *T. colubriformis* L3 have been recovered in the absence of rainfall  
310 (O'Connor et al., 2008; Wang et al., 2014). Therefore a 4- and 7-day trailing  
311 cumulative P/E rule was applied to *H. contortus* and *T. circumcincta* respectively to  
312 characterise sub-optimal conditions, extrapolated from the observations of O'Connor  
313 et al. (2008) and Khadijah et al. (2013b) on the effect of cumulative P/E on FMC and  
314 development success. Sub-optimal days were defined as days where total  
315 precipitation < 2mm, and trailing cumulative P/E ≥ 1. The species specific trailing  
316 cumulative P/E values for migration are referred to as P/E<sub>4</sub> for *H. contortus* and  
317 P/E<sub>7</sub> for *T. circumcincta*.

318 An estimated horizontal migration rate of 0.051 for *H. contortus* under sub-optimal  
319 moisture conditions was derived from observations by O'Connor et al. (2008), where  
320 30% of L3 migrated out of faeces within a 7 day period in the absence of rain but  
321 following a period of simulated rainfall in the preceding 7 days. This is consistent with  
322 the observed mean migration rate of 0.057 (S.D. 0.027) for *H. contortus* maintained  
323 under high relative humidity of 98% and FMC of ~60% (Wang et al., 2014).

324 To estimate the corresponding rate for *T. circumcincta* the instantaneous daily  
325 migration rate of *H. contortus* L3 estimated from data provided by Wang et al. (2014)  
326 was compared with the instantaneous daily migration rate of *T. circumcincta*  
327 estimated from an unpublished experiment conducted concurrently with the  
328 experiment of Wang et al. (2014) and using identical methods. These experiments



329 show that under optimal moisture conditions the instantaneous daily migration rate of  
330 *T. circumcincta* is 49% that of *H. contortus*. Ninety-nine percent (S.D. 0.4) of *H.*  
331 *contortus* L3 had migrated out of faeces within 7 days (Wang et al. 2014), compared  
332 with 91% (S.D. 6.8) of *T. circumcincta* L3, giving instantaneous daily horizontal  
333 migration rates of 0.71 and 0.35 respectively. Thus the estimated instantaneous daily  
334 migration rate for *T. circumcincta* where moisture is sub-optimal is  $0.051 \times 0.49 =$   
335 0.025 (Table 2).

336 Finally, a horizontal migration rate of 0 was applied for all GIN species on moisture-  
337 limited days where  $P/E < 1$  and total precipitation  $< 2\text{mm}$ .

338

### 339 **2.2.3 Migration between soil and herbage**

340 Crofton (1948) observed seasonal patterns in the vertical migration of  
341 *Trichostrongylus retortaeformis* L3 on pasture where fewer L3 were recovered from  
342 the upper herbage layer in winter than in summer. It is likely that interacting climatic  
343 and other abiotic variables including temperature, moisture, biomass composition  
344 and light drive this seasonality (Amaradasa et al., 2010; Callinan and Westcott,  
345 1986; Crofton, 1948; Dusenbery, 1989; Ogbourne, 1973; Rees, 1950; Saunders et  
346 al., 2000; Silangwa and Todd, 1964; van Dijk et al., 2009). However, the majority of  
347 studies have sampled only a superficial layer of soil (e.g. Crofton, 1948; Rees, 1950)  
348 and therefore could underestimate the proportion of pasture L3 in soil relative to L3  
349 on herbage. Mesocosm experiments (e.g. Callinan and Westcott, 1986; Knapp-  
350 Lewitzke et al. in preparation) offer an alternative to ensure more complete sampling  
351 of L3 in soil. The temperature-dependent proportion of trichostrongylid L3 expected  
352 on herbage and in soil was estimated by fitting a second order polynomial regression

353 to the log transformed proportion of total *Teladorsagia* and *Trichostrongylus* spp. L3  
354 recovered from herbage (Callinan and Westcott, 1986). In the absence of suitable  
355 species-specific data, the same estimate was used for all trichostrongylid GIN  
356 species, subject to validation. The observation at 20°C was omitted from analysis as  
357 the decrease in L3 recovered from herbage was inconsistent with observations of L3  
358 availability on pasture where the percentage of L3 recovered from herbage tends to  
359 increase with increasing mean soil temperature between approximately 8-22°C  
360 (Callinan, 1979, 1978a, 1978b).

361

### 362 **2.3 Model validation**

363 Laboratory observations on the development success of *T. circumcincta* and *O.*  
364 *ostertagi* (Rossanigo and Gruner, 1995) and field observations on the development  
365 time and development success of *H. contortus* (Rose, 1963) and *O. ostertagi* (Rose,  
366 1961) in a temperate region were used to validate model predictions of development  
367 and survival in faeces. Rossanigo and Gruner (1995) recorded the development  
368 success of various GIN species in faeces incubated at a range of constant  
369 temperature while maintaining optimal moisture conditions. Rose (1963, 1961)  
370 recovered L3 from faeces containing eggs that had been deposited at monthly  
371 intervals on pasture in South East England where the GINs were exposed to variable  
372 temperatures and moisture conditions. The model was initialised with 100 eggs and  
373 simulations were run using either the constant temperatures tested (Rossanigo and  
374 Gruner, 1995) or time series of daily air temperatures obtained for the study location.  
375 Meteorological data from Wisley weather station (Ordnance Survey grid reference:  
376 TQ062579), approximately 10km from the Central Veterinary Laboratories,

377 Weybridge, where Rose's (1963) observations were made, were obtained from the  
378 British Atmospheric Data Centre (badc.nerc.ac.uk). The same data were not  
379 available for the time period of observations made by Rose (1961). Therefore,  
380 meteorological data were obtained from the E-OBS gridded dataset (lat/lon:  
381 51.355°N, -0.496°E; Haylock et al., 2008). Potential evaporation (mm/day) was  
382 estimated from mean air temperatures using the Hamon method (Xu and Singh,  
383 2001). Two simulations were run for each monthly deposit using mean daily  
384 temperature and linear fluctuations between minimum and maximum daily  
385 temperature to determine whether mean air temperature is sufficient to predict  
386 development when temperatures are close to the minimum threshold for  
387 development. Horizontal migration,  $m_1$ , was set to 0 to prevent migration out of  
388 faeces.

389 Field observations of *H. contortus* and other trichostrongylid  
390 (*Trichostrongylus/Teladorsagia* spp.) L3 over winter on naturally contaminated  
391 pasture in the absence of continued grazing by livestock (Wilkie et al. submitted)  
392 were used to validate the predicted dynamics of *H. contortus* and *T. circumcincta* L3  
393 availability on herbage. Temperature data for the observation period were obtained  
394 from Yeovilton weather station (Ordnance survey grid reference: ST549231),  
395 approximately 60km from the farm where observations were made from the British  
396 Atmospheric Data Centre. The initial number of L3 recovered from herbage at the  
397 start of the observation period and the vertical migration parameter,  $m_2$ , were used to  
398 estimate the corresponding initial number of L3 expected in soil. All other initial  
399 values were set to 0. For each simulation the daily number of L3 on herbage is a  
400 product of the daily number of L3 on pasture,  $L3p$ , and the vertical migration  
401 parameter,  $m_2$ .

402 To determine whether the additional complexity of the pasture component of the  
403 GLOWORM-FL model was justified, simulations using an existing model for *H.*  
404 *contortus* (Smith, 1990) were also validated using Wilkie's (submitted) data as  
405 described above.

406 For each validation dataset and corresponding simulations, model fit was assessed  
407 using the residual sum of squares (RSS; Mayer and Butler, 1993) and linear  
408 regression through the origin. An intercept of 0 and slope of 1 would indicate perfect  
409 correspondence between model output and observations, therefore a regression  
410 through the origin with a slope that is not statistically significantly different from 1  
411 indicates a good model fit. It is assumed that the slope is not statistically significantly  
412 different from 1 if the 95% confidence interval (estimated as the coefficient  $\pm(2 \times$   
413 standard error of the coefficient) includes 1.

#### 414 **2.4 Climate change simulations**

415 The validated model was run using mean daily temperature and total daily  
416 precipitation data from the atmospheric dataset provided by the Coupled Model  
417 Intercomparison Project Phase 5 (CMIP5; Taylor et al., 2012) to predict the potential  
418 impact of current climate change predictions on the seasonal availability of L3 on  
419 pasture (infection pressure). Simulations ran for 30-year time periods using either  
420 historical climatic data for the period 01/12/1969-30/11/1999 (representative of  
421 current climate) or a high emissions scenario (Representative Concentration  
422 Pathway 8.5; RCP8.5) for the period 01/12/2070-30/11/2100, from the HadGEM2-ES  
423 model output (ensemble r1i1p1) developed and run by the Met Office Hadley Centre  
424 (Collins et al., 2011; Martin et al., 2011). Characteristics of the RCP8.5 scenario  
425 include high greenhouse gas emissions, a high rate of population growth, a

426 dependence on fossil fuel and global CO<sub>2</sub> concentrations of ~950ppm by 2100 (van  
427 Vuuren et al., 2011). For comparison, record CO<sub>2</sub> concentrations of over 400ppm  
428 were observed at the Mauna Loa observatory in May 2014 (Tans, 2014).

429 Time series of mean daily air temperature and total daily precipitation were extracted  
430 for a grid cell in North Somerset in South West England, UK. This area is of  
431 particular interest as recent climate change has been associated with an increase in  
432 diagnoses of parasitic gastroenteritis in the region (van Dijk et al., 2008).

433 One hundred new eggs ( $E_{new}$ ) were added daily to simulate a scenario of continuous  
434 grazing and host infection without making assumptions about management or  
435 seasonal changes in intensity of infection/nematode egg output. The first year of  
436 simulation was discarded as L3 accumulated on pasture throughout the first year.  
437 Output is presented as annual time series of daily mean numbers of L3 on pasture,  
438 calculated using the remaining 29-year output disaggregated into annual time series.

439 The area under curve (AUC) was calculated for each year using a trapezoid function  
440 in R to estimate the annual infection pressure under the historical and future climate  
441 scenarios. Wilcoxon rank sum tests were used to compare scenarios for each GIN  
442 species (Figure 5).

443

## 444 **3. Results**

### 445 **3.1 Model validation**

446 Overall the model was able to reproduce the observed development times,  
447 development success, and dynamics of L3 on pasture (Table 3), demonstrating the  
448 potential for a generic framework such as GLOWORM-FL to be adapted to suit  
449 different GIN and host species.

450 A slope marginally greater than 1 suggested that there was a tendency to  
451 under-predict the development success of *T. circumcincta* and *O. ostertagi* at  
452 constant temperatures between 5°C and 35°C compared with laboratory  
453 observations (Table 3).

454 The model performed well when tested against field observations of *H. contortus* and  
455 *O. ostertagi* development and survival in faeces (Table 3) demonstrating that the  
456 models, parameterised using laboratory data, transferred well onto conditions  
457 observed in the field. The range of mean air temperatures during the observation  
458 periods was -3.9 to 24.7°C (Rose, 1963) and -2.5 to 23.2°C (Rose, 1961). The range  
459 of total daily precipitation during the observation periods was 0-31.2mm (Rose,  
460 1963) and 0-23.2mm (Rose, 1961).

461 The predicted dynamics and numbers of *H. contortus* and *T. circumcincta* L3 on  
462 pasture fitted observations well, replicating the initial decrease in L3 density on  
463 herbage followed by an increase, despite no further contamination of pasture (Figure  
464 6; Table 3). This seasonal variability in L3 on herbage can be explained by the  
465 vertical migration of L3 between soil and herbage; a model using only a temperature-  
466 dependent mortality rate and not considering movement between the soil and

467 herbage was not able to replicate these dynamics. The range of mean daily air  
468 temperatures during the observation period was -3.05 to 13.7°C.

469 The GLOWORM-FL model, validated using observed numbers of *H. contortus* on  
470 pasture, outperformed a previously published, less complex model (Table 3; Figure  
471 6).

472 The performance of models using minimum and maximum or mean daily  
473 temperatures varied dependent on the validation dataset (Table 3) and in most  
474 cases both models gave similar output. However, models using minimum and  
475 maximum daily temperatures performed poorly against observation of *O. ostertagi*  
476 development in the field (Table 3). Therefore mean temperatures were used for all  
477 subsequent simulations.

### 478 **3.2 Climate change simulations**

479 At the chosen test location in South West England, UK, the HadGEM-ES model  
480 predicts warmer wetter winters and warmer, drier summers during 2070-2100 under  
481 the RCP8.5 high emissions scenario, compared with the historical period of 1969-  
482 1999. A mean increase in mean air temperature of 4.57°C (S.D. 1.91°C) is predicted  
483 by 2070-2100. The increase is greatest during the summer months with a maximum  
484 of 8.65°C increase predicted during July and a minimum of 1.49°C increase  
485 predicted during March. A mean decrease of 0.03mm (S.D. 1.09mm) in mean daily  
486 rainfall is predicted under the RCP8.5 scenario, with an increase of up to 3.02mm  
487 during the winter period and a decrease of up to 3.06mm during the summer period.  
488 The change in seasonal temperatures and rainfall resulted in an increase in  
489 predicted development rate throughout the year whereas mortality rates decreased  
490 during the winter and increased during the summer. The pattern of moisture-

491 limitation on development success and horizontal migration of *T. circumcincta* and *H.*  
492 *contortus* was similar under both scenarios. Although the patterns of change in life-  
493 history parameters were similar, the magnitude of change was species dependent,  
494 resulting in differing seasonal patterns of L3 on pasture.

495 There was a significant predicted increase in annual infection pressure for both *H.*  
496 *contortus* ( $W=47$ ,  $p<0.001$ ) and *T. circumcincta* ( $W=95$ ,  $p<0.001$ ; Figure 5) under the  
497 RCP8.5 scenario compared with historical climatic data. Mean air temperature was  
498 regularly higher than the predicted lower threshold for development of *T.*

499 *circumcincta* ( $4.46^{\circ}\text{C}$ ) when both historical and RCP8.5 data were used and  
500 development was possible year round. However, the number of days where  
501 development was possible increased from 328 to 360 (the HadGEM-ES model was  
502 run on a 360 day year and therefore 360 represents the entire year). The increase in  
503 temperatures predicted under RCP8.5 therefore resulted in increased development  
504 rates year-round for *T. circumcincta*. In contrast, very little *H. contortus* development  
505 is completed over winter when using historical climatic data as the mean air  
506 temperature rarely rises above the predicted threshold for development of  $9.17^{\circ}\text{C}$ .

507 Therefore, the increase in temperatures predicted under RCP8.5 not only results in  
508 an increase in development rate but also a lengthening of the season during which  
509 development is possible. The period during which mean daily temperatures  
510 exceeded the development threshold for *H. contortus* was extended by 3.3 months  
511 from 188 days between March and September under historical climate to 258 days  
512 between February and December under RCP8.5. A corresponding decrease in  
513 mortality rates during the winter results in an overall increase in infection pressure  
514 over the winter period, extending into early summer. Further increases in  
515 temperatures during the summer result in increased mortality which offsets the



516 increased development rate and results in a decrease in the number of L3 on  
517 pasture, below numbers predicted using historical data (Figure 7).

518 A similar pattern of summer mortality is predicted for *O. ostertagi* (Figure 7). When  
519 using historical climatic data, there is a small increase in L3 on pasture during the  
520 spring as temperatures exceed the predicted threshold for development of 7.44°C,  
521 but the large number of L3 predicted on pasture is fairly consistent throughout the  
522 year due to low mortality rates. However, there is a significant decrease in annual  
523 infection pressure when using the RCP8.5 climatic data, compared with predictions  
524 using historical data ( $W=95$ ,  $p<0.001$ ). The period during which mean daily  
525 temperatures exceeded the development threshold for *O. ostertagi* was extended by  
526 4.6 months from 216 days between March and October under historical climate to  
527 347 days throughout the year under RCP8.5, but this is offset by significant  
528 increases in mortality rates between May and November depleting the reservoir of  
529 L3 in faeces and soil.

530

531

#### 532 **4. Discussion**

533 The model presented here consolidates advances in our understanding of the  
534 ecology and behaviour of gastrointestinal nematode free-living stages with the  
535 numerous existing models developed to simulate the population dynamics of GINs in  
536 ruminant livestock. Previous models were species-specific (Grenfell et al., 1986;  
537 Smith, 1990), restricted to livestock ruminants (Learmount et al., 2006), and  
538 constrained by the available data at the time they were developed. GLOWORM-FL  
539 builds significantly on these models to incorporate the active movement of GINs  
540 between substrates, and substrate-specific mortality rates, in addition to explicitly  
541 climate-dependent life-history parameters. Comparison of the output of GLOWORM-  
542 FL with an example of a preceding model that did not consider nematode behaviour  
543 with field observations confirmed that the additional complexity significantly improved  
544 model predictive performance. This is probably because the majority of L3 are  
545 sequestered in the soil at any one time (Callinan and Westcott, 1986; Silangwa and  
546 Todd, 1964; van Dijk and Morgan, 2011), and emerge onto herbage when climatic  
547 conditions allow. Therefore, soil should not be overlooked as a significant source of  
548 infection, acting as a reservoir for L3 that can recolonize herbage (van Dijk and  
549 Morgan, 2011). For the same reason, absence of L3 on herbage should not be  
550 interpreted as evidence for absence of GINs.

551 An additional motivation for the development of the model was that characteristics  
552 relevant to the epidemiology of GINs are similar between different livestock systems  
553 and wild ruminants (Rose et al., 2014) and there is evidence of transmission where  
554 livestock and wildlife meet or share ranges (Chintoan-Uta et al., 2014; Morgan et al.,  
555 2007). There is therefore a need for a common framework that can be applied to a  
556 range of GIN and host species. The GLOWORM-FL model framework was

557 parameterised and successfully validated using data available in the literature for  
558 three GIN species that are economically important parasites of livestock worldwide  
559 but also infect free-living ruminants (Morgan et al., 2005). Due to their economic  
560 importance, research on these species spans decades, thus providing sufficient data  
561 for parameter estimation. There were some gaps in the available data and therefore  
562 these species also provided an opportunity to demonstrate how the model can be  
563 successfully adapted by drawing on similarities between GIN species and robust  
564 validation exercises.

565 GINs are a global constraint on livestock production (Nieuwhof and Bishop, 2007;  
566 Perry and Grace, 2009). The increasing prevalence of anthelmintic resistance  
567 worldwide (Kaplan and Vidyashankar, 2012) and the threat of altered seasonal  
568 patterns of transmission due to climate change (Gethings et al. submitted; Molnár et  
569 al., 2013; van Dijk and Morgan, 2010; van Dijk et al., 2008) necessitate the  
570 development of alternative control strategies (Krecek and Waller, 2006). It may be  
571 possible to control the magnitude of exposure to GINs and therefore intensity of  
572 infection and production losses, for example by altering management practices to  
573 avoid grazing during periods of high risk or targeting treatments according to risk of  
574 exposure or suitability for development of free-living stages. To do this,  
575 understanding the population ecology of GINs and predicting the seasonal dynamics  
576 of infection pressure is fundamental.

577 GLOWORM-FL provides a tool to aid in the development of climate-based GIN  
578 control methods. The model can be used to track pasture contamination and  
579 evaluate the resultant climate-dependent infection pressure under a range of  
580 management and climate scenarios. Here, its use is demonstrated using climatic  
581 data representative of recent historical climate and climate expected under the

582 IPCC's RCP8.5 scenario. Climate-driven changes in the seasonal availability of L3  
583 on pasture are likely to become increasingly important in the dynamics of GIN  
584 infection, particularly where host behaviour or farm management is slow to adapt in  
585 response to the change. In some cases, this may lead to an increase in disease (van  
586 Dijk et al., 2008) whereas under different circumstances climate-driven changes may  
587 decrease exposure to infection, as has apparently been the case for *Nematodirus*  
588 *battus* infections in lambs in some parts of the UK (Gethings et al. submitted). A  
589 better understanding of the seasonal dynamics of infection pressure will be key to  
590 the future of sustainable GIN control in livestock and could also benefit the  
591 management and conservation of wild ruminants.

592 Using historical climatic data, large numbers of *O. ostertagi* L3 were predicted year  
593 round on pasture due to low mortality rates over winter and a turnover of L3 between  
594 April and November when development rates increase and compensate for losses  
595 due to the increased mortality rate. This suggests that the observed patterns of  
596 ostertagiasis in calves in Europe (Williams et al., 1993), where peak worm burdens  
597 are observed towards the end of the grazing season, are driven by cumulative  
598 exposure to L3 on pasture and management or host factors as opposed to seasonal  
599 variability in infection pressure (Höglund et al., 2013; Roberts and Grenfell, 1992).

600 Under the RCP8.5 climate scenario and a constant input of eggs, a decrease in *O.*  
601 *ostertagi* infection pressure is predicted throughout the year due to significant  
602 increases in predicted mortality rates depleting the reservoirs of infective stages in  
603 faeces and on pasture. Although a reduction in the magnitude of exposure to  
604 infective stages is favourable, there may also be an adverse impact on the  
605 development of immunity through reduced exposure to L3 (Ploeger et al., 1995).  
606 However, since the epidemiology of *O. ostertagi* infection is largely driven by

607 management and host factors (Höglund et al., 2013; Roberts and Grenfell, 1992),  
608 altered management strategies in response to climate change may negate the  
609 change in seasonal availability of L3 on pasture predicted here.

610 The seasonal incidence of *H. contortus* infection is primarily climate-driven, and the  
611 pattern predicted here for South West England using the historical climatic data  
612 broadly mirrors the seasonal diagnoses of haemonchosis in sheep in the region (van  
613 Dijk et al., 2008). The implications of this for the control of haemonchosis in livestock  
614 are that predicted changes in L3 on pasture are likely to result in similar changes in  
615 the seasonal incidence of haemonchosis. The predicted pattern of infection pressure  
616 for *T. circumcincta* when using historical climatic data also reflects patterns of  
617 seasonal diagnoses (van Dijk et al., 2008), indicating a degree of climate-  
618 dependence in the epidemiology of *T. circumcincta* infection in sheep in South West  
619 England.

620 An increase in temperature during the winter months was predicted, resulting in an  
621 increase in infection pressure for both *H. contortus* and *T. circumcincta* due to a  
622 corresponding increase in development rates. Development of *T. circumcincta* is  
623 possible throughout the year in South West England. In temperate regions *H.*  
624 *contortus* survival on pasture over winter is poor and there is very little development  
625 of eggs deposited on pasture as temperatures fall below the development threshold.  
626 However, *H. contortus* is able to survive the winter period as arrested larvae within  
627 the host (Waller et al., 2004). The increase in temperatures predicted here for South  
628 West England, extends the period where the development of *H. contortus* is possible  
629 and could have significant short- and long-term impacts on the epidemiology of *H.*  
630 *contortus* in temperate regions. In the short-term, the increase in infection pressure  
631 throughout the year will result in year-round transmission. In the long-term, *H.*

632 *contortus* may adapt in response to the reduced selection pressure for arrested  
633 development (hypobiosis) in the host, potentially resulting in a decreased propensity  
634 to arrest. Using a series of mathematical models, Dobson and Hudson (1992)  
635 showed that hypobiosis decreases the basic reproductive rate ( $R_0$ ) of trichostrongylid  
636 nematodes. Gaba and Groubière (2008) built on the work of Dobson and Hudson  
637 and further demonstrated the potentially destabilising effect of hypobiosis on GIN  
638 population dynamics as the mortality rate of the free-living stages is decreased.  
639 Therefore, hypobiosis would not be favoured when climatic conditions render it  
640 unnecessary.

641 The predicted increase in the availability of *H. contortus* and *T. circumcincta* L3 on  
642 pasture during the spring under the RCP8.5 scenario is a concern as this coincides  
643 with peak lambing/kidding and peak parturition in many wild ruminants. Therefore,  
644 naïve individuals may experience a much greater challenge early in the grazing  
645 season under this scenario of climate change. Ewes experiencing a breakdown in  
646 immunity to gastrointestinal nematodes during pregnancy and lactation (Houdijk et  
647 al., 2001) will also experience a greater challenge. For example, an increase in *H.*  
648 *contortus* infection pressure during the spring could result in more acute  
649 haemonchosis in naïve individuals and increased pasture contamination early in the  
650 grazing season. These effects may be magnified by management and host factors.  
651 The current model considers a scenario of constant pasture contamination.  
652 However, pasture contamination may increase during the spring reproductive period  
653 due to the periparturient rise (PPR) in faecal egg counts observed in reproducing  
654 animals, which is due to the maturation of hypobiotic larvae resulting from a complex  
655 of factors thought to result in a reduction in immunity during pregnancy and lactation  
656 (Falzon et al., 2013; Gibbs and Barger, 1986; Houdijk et al., 2001).

657 A decrease in infection pressure is predicted during the summer months for all  
658 species tested as a result of a trade-off between increased development and  
659 increased mortality rates. As discussed for *O. ostertagi*, reduced exposure to L3 may  
660 impact on the development of immunity. However, these reductions may be  
661 dampened by increasing worm burdens in the host and therefore increased pasture  
662 contamination throughout the grazing season. Furthermore, there is potential for  
663 parasite adaptation in response to decreased transmission and the impact this has  
664 on host immunity. For example, nematode fecundity may be negatively associated  
665 with host immune response as suggested by the negative correlation between adult  
666 *T. circumcincta* length and immune response (mucosal and serum IgA against L3  
667 and L4), and the positive correlation between worm length and number of eggs in  
668 (nematode) utero in artificially infected lambs (Stear et al., 1995). Integrating the  
669 GLOWORM-FL framework with models of the parasitic stages, host immunity (e.g.  
670 Grenfell et al., 1987) and parasite adaptation will allow the impact of changes in the  
671 seasonal exposure to L3 on the potential pathogenesis of infection and subsequent  
672 population dynamics of parasites on pasture to be explored.

673 Validation was successful for all three species tested, not only validating the model  
674 structure but also demonstrating that gaps in parameter estimates can be addressed  
675 using data from other species, that parameter estimates derived from laboratory  
676 observations perform well under conditions experienced in the field, and that data  
677 obtained from the nearest weather station can be used in the absence of local  
678 meteorological observations.

679 In some cases, linear regression of observations against model predictions was  
680 significant, but the slope of the regression was marginally different from 1, indicating  
681 systematic bias in the output. However, in most cases the error was within the range

682 expected from factors such as trait variation (Troell et al., 2006; van Dijk and  
683 Morgan, 2010), measurement error (Persson, 1974b) and uncertainty arising from  
684 model structure.

685 Simulations using mean daily temperature data outperformed those using minimum  
686 and maximum data. Minimum and maximum daily temperature data were used to  
687 test whether fluctuations above and below the development thresholds were  
688 important in predicting the population dynamics free-living GINs. At certain times of  
689 year the mean temperature may be above the threshold for development, but if the  
690 minimum temperature falls below the threshold there is potential to over-predict the  
691 development rate using only mean temperatures. Conversely, if the mean  
692 temperature is below the threshold but the maximum falls above the threshold, then  
693 models may fail to predict development at all. It was therefore surprising that  
694 allowing temperatures to fluctuate between the minimum and maximum daily values  
695 did not improve model performance.

696 Discrepancies between meteorological observations and microclimatic conditions  
697 may account for the superior performance of simulations using mean daily air  
698 temperatures. Recent studies have demonstrated the importance of microclimatic  
699 factors in determining GIN abundance under controlled conditions (Khadijah et al.,  
700 2013b; O'Connor et al., 2008; Wang et al., 2014). In the field, temperature and  
701 moisture fluctuations in faeces may be buffered by the soil beneath and surrounding  
702 herbage. This buffering effect may also explain discrepancies between model  
703 predictions and observations e.g. the model underestimated the time to development  
704 of *H. contortus* L3 observed by Rose (1963) during April and October regardless of  
705 whether minimum-maximum or mean temperatures were used. Soil temperature  
706 data were not available for use in these validation exercises but may better reflect



707 the microclimate around faeces. Where possible, further validation should also  
708 include soil temperature.

709 The model was validated using observations made in a temperate region, with  
710 temperatures ranging between - 4°C and 25°C. However, temperatures of up to  
711 39.2°C were predicted under the RCP8.5 scenario and therefore some simulations  
712 projected beyond the range of the conditions used for validation. *H. contortus* and *T.*  
713 *circumcincta* simulations using climatic data for the RCP8.5 scenario showed that  
714 high temperatures may result in counter-intuitive decreases in the availability of L3  
715 due to a trade-off between increased mortality and development rates. Uncertainty in  
716 climate change simulations due to projections outside of the range of observed data  
717 could be reduced by repeating validation using data from regions with current  
718 climatic conditions similar to those predicted under the chosen climate change  
719 scenario.

720 The development of parsimonious mechanistic models is inevitably a compromise  
721 between biological realism, complexity and the availability of data for parameter  
722 estimation. Here we have used observations on mortality of L3 in water to estimate  
723 mortality rates for L3 in soil. Although these estimates are similar to published  
724 observations of L3 mortality in soil (van Dijk and Morgan, 2011) and the model was  
725 able to predict the survival of L3 on pasture in a temperate region well, site-specific  
726 and temporal variations in soil conditions such as moisture content, pH and the  
727 presence of nematophagous fungi may affect the observed mortality rates and  
728 increase model uncertainty. Variations in faecal moisture and structural integrity of  
729 faeces in the field may also affect the population dynamics of GINs. Diarrhoea is  
730 commonly associated with infection by GINs such as *T. circumcincta* and *O.*  
731 *ostertagi* but can also be attributed to a range of other causative agents such as diet

732 and other gastrointestinal infections. As such it is difficult to characterise faecal  
733 consistency for inclusion in mechanistic models, but this potential source of variation  
734 should be noted, especially in the context of development success and horizontal  
735 migration of L3 between faeces and pasture.

## 736 **5. Conclusion**

737 A general model framework was developed to simulate the climate-driven population  
738 dynamics of the free-living stages of trichostrongylid GINs. Simulations using  
739 historical and future climatic data predicted significant changes in seasonal and  
740 annual infection pressure in the absence of host management, including a surprising  
741 decrease in infection pressure for *O. ostertagi*. Integration with management data,  
742 host behaviour and models developed to simulate the parasitic stages of these  
743 species, will enable the evaluation of GIN control options under a range of climate  
744 scenarios to identify long-term sustainable strategies.

745

746

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765

766

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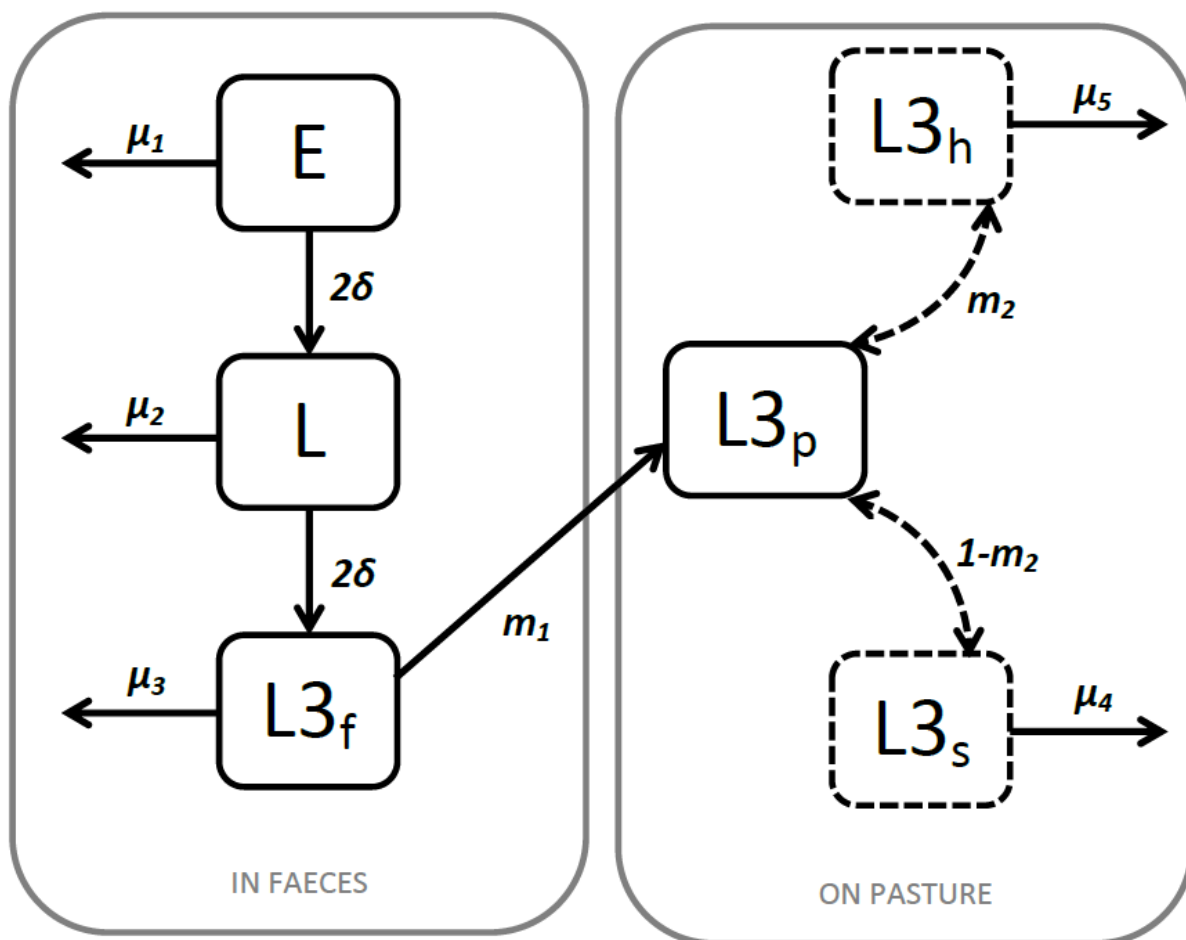


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1042 Figure 1. Conceptual diagram of the GLOWORM-FL model framework. Parameter  
1043 (lower case) and state variable (upper case) definitions are given in Table 1.

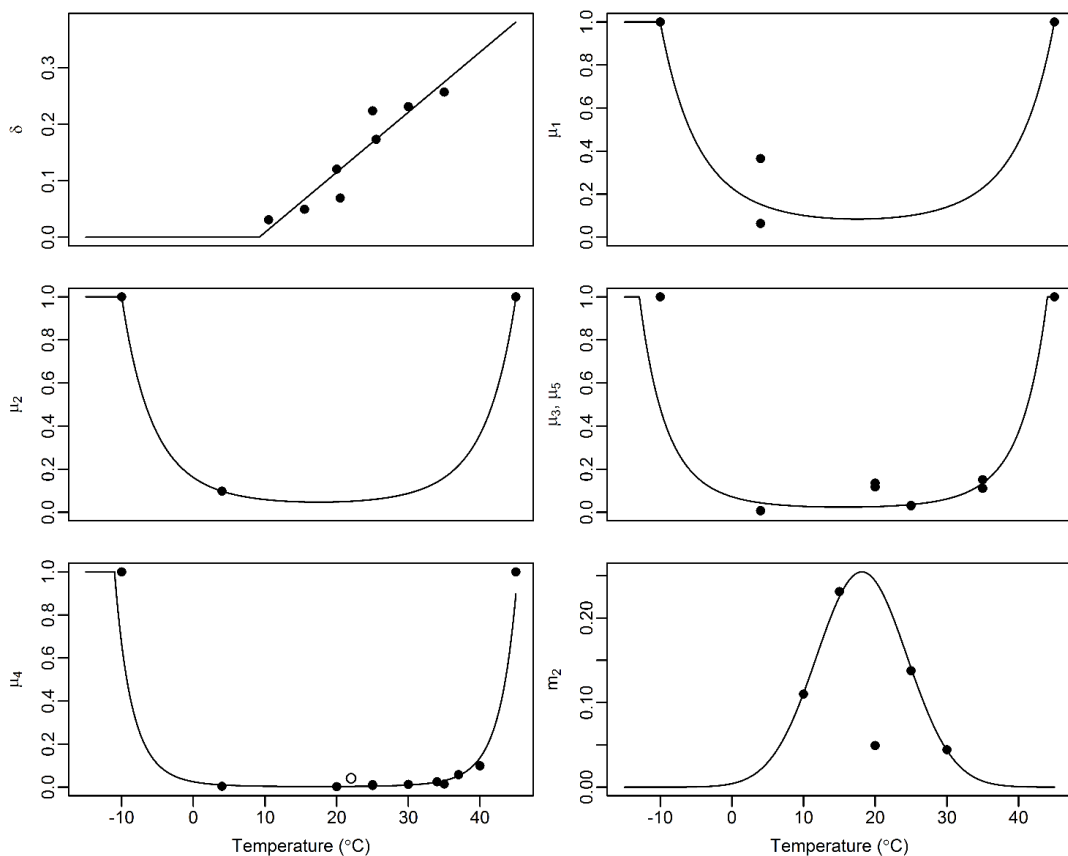


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1047 Figure 2. Estimates of temperature-dependent life-history parameters for *H.*  
 1048 *contortus* (lines) based on analysis of data in the literature (closed circles).  
 1049 Parameter definitions are given in Table 1. Statistical output for linear models is  
 1050 provided in Table 2. Mortality of L3 in soil ( $\mu_4$ ) was estimated from observations of L3  
 1051 mortality in water. A point estimate of L3 mortality in desiccated soil at 20-24°C (van  
 1052 Dijk and Morgan, 2011) is superimposed (open circle) for comparison. Data were not  
 1053 available to estimate the mortality of L3 on herbage ( $\mu_5$ ), which was therefore  
 1054 estimated using the mortality rate of L3 in faeces ( $\mu_3$ ). The data point at 20 degrees  
 1055 was omitted from analysis of the vertical migration parameter ( $m_2$ ) but is shown here.  
 1056 A minimum threshold for development of 9.17°C is predicted.

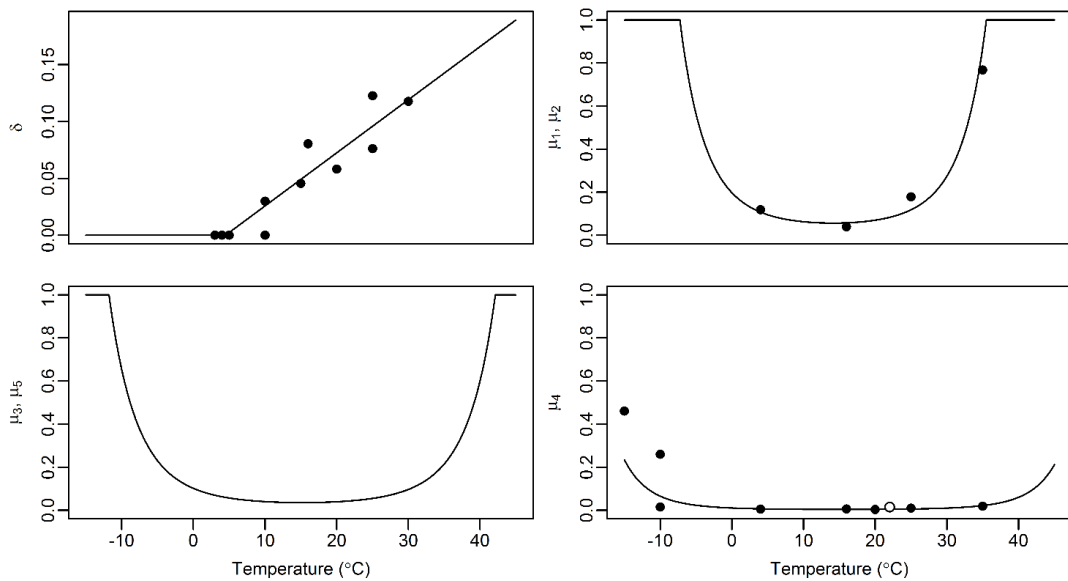


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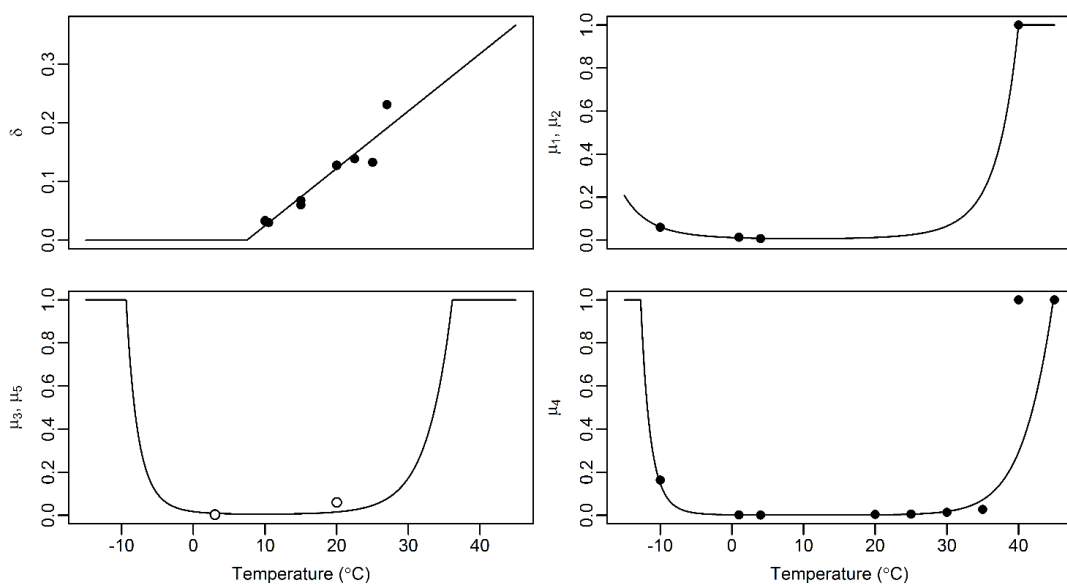
1060 Figure 3. Estimates of temperature-dependent life-history parameters for *T.*  
 1061 *circumcincta* (lines) based on analysis of data provided in the literature (closed  
 1062 circles). Parameter definitions are given in Table 1. Statistical output for linear  
 1063 models is provided in Table 2. Mortality of L3 in soil ( $\mu_4$ ) was estimated from  
 1064 observations of L3 mortality in water. A point estimate of L3 mortality in desiccated  
 1065 soil at 20-24°C (van Dijk and Morgan, 2011) is superimposed (open circle) for  
 1066 comparison. Data points are not shown for the mortality rates of L3 in faeces ( $\mu_3$ )  
 1067 and on herbage ( $\mu_5$ ) as no data were available to directly estimate these parameters.  
 1068 These rates were therefore estimated from the mortality rate of L3 in soil ( $\mu_4$ ). A  
 1069 minimum threshold for development of 4.46°C is predicted. The vertical migration  
 1070 parameter is as shown in Figure 2.







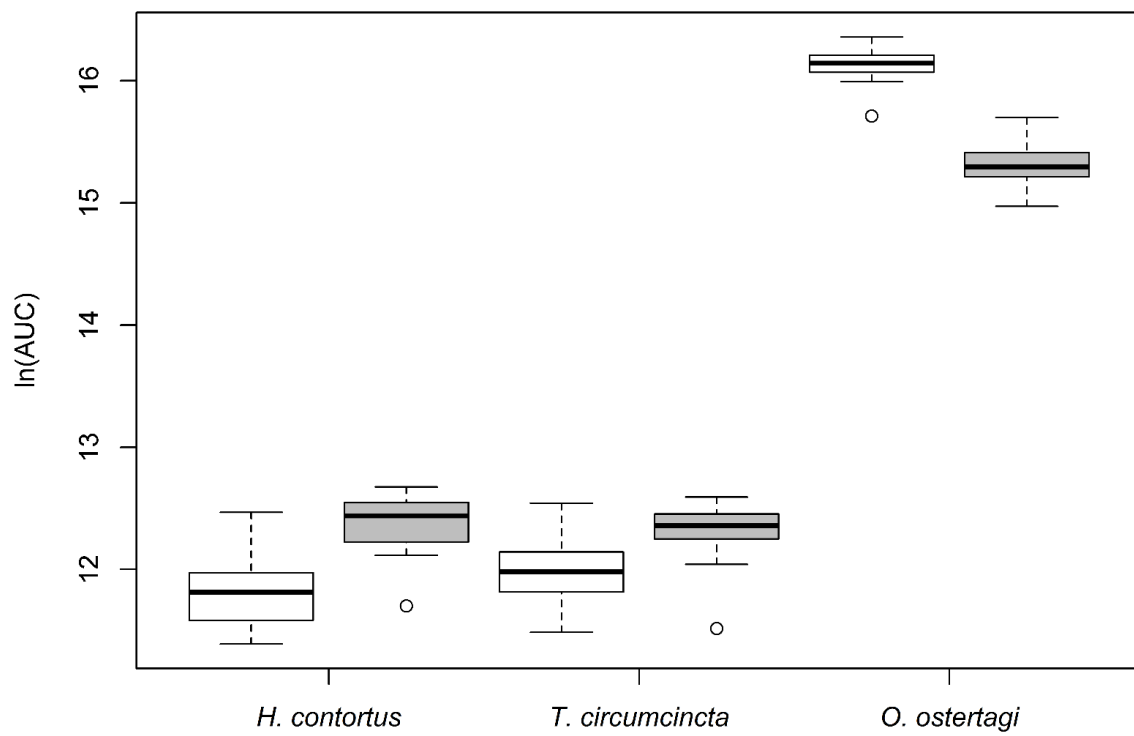
1073 Figure 4. Estimates of temperature-dependent life-history parameters for *O. ostertagi*  
 1074 (lines) based on analysis of data provided in the literature (closed circles). Parameter  
 1075 definitions are given in Table 1. Statistical output for linear models is provided in  
 1076 Table 2. Data points are not shown for the mortality rates of L3 in faeces ( $\mu_3$ ) and on  
 1077 herbage ( $\mu_5$ ) as no data were available to directly estimate these parameters. These  
 1078 rates were therefore estimated from the mortality rate of L3 in soil ( $\mu_4$ ). Estimates of  
 1079 mortality in faeces ( $\mu_3$ ) based on analysis of observations on *Cooperia oncophora*  
 1080 and *O. ostertagi* mixed infections (open circles) are superimposed for comparison  
 1081 (Persson, 1974a). A minimum threshold for development of 7.44°C is predicted. The  
 1082 vertical migration parameter is as shown in Figure 2.



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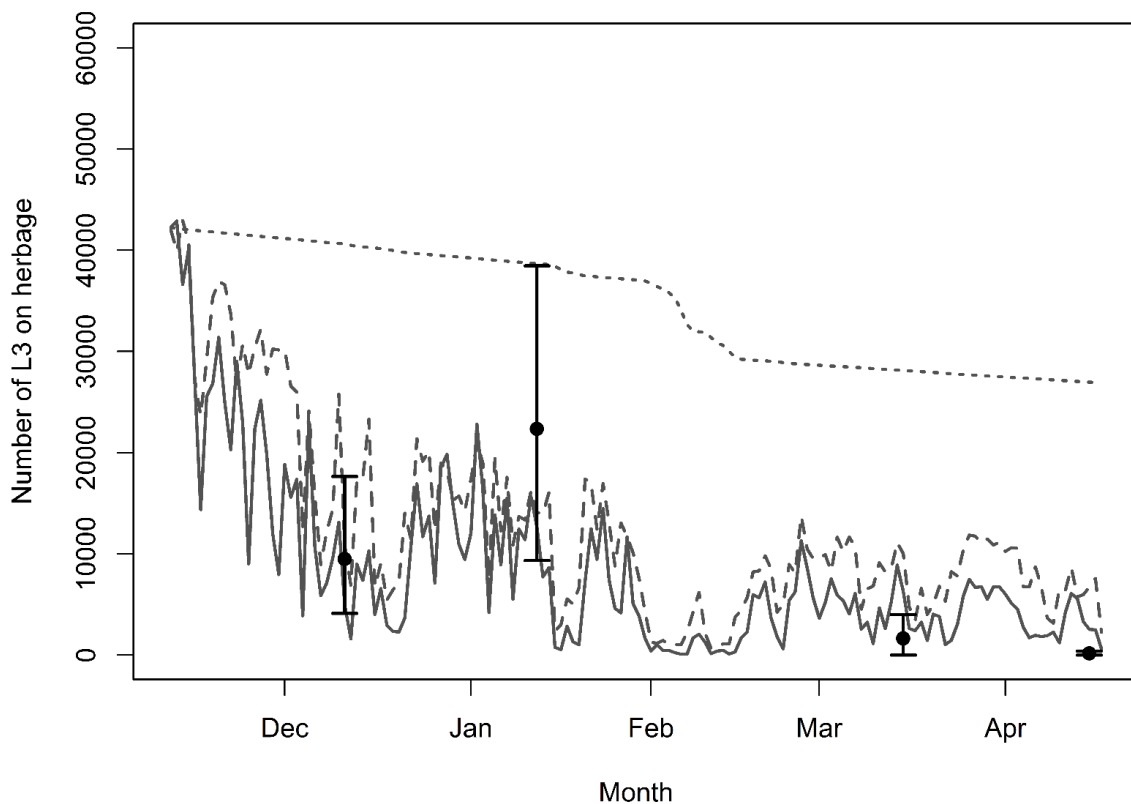
1085 Figure 5. Estimated annual AUC (Area Under the Curve) of the predicted numbers of  
1086 L3 on pasture for *H. contortus*, *T. circumcincta* and *O. ostertagi* when using  
1087 historical climatic data for the period 1969-1999 (white) and climatic data based on  
1088 the RCP8.5 high emissions climate change scenario for the period 2070-2100 (grey).



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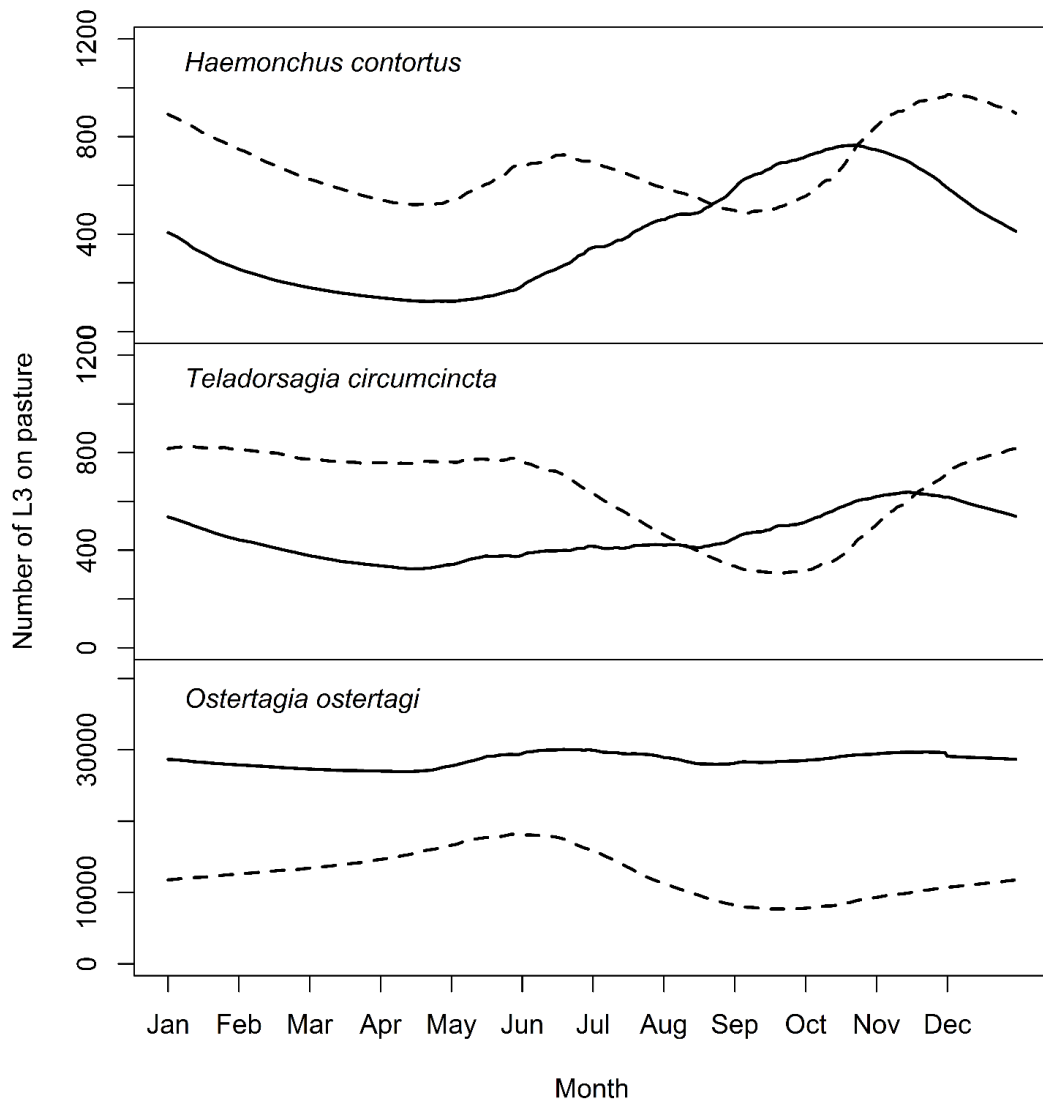
1091 Figure 6. The number of *H. contortus* L3 on herbage (L3 kg DM-1) predicted using  
1092 the GLOWORM-FL model and mean daily air temperature (dashed line) or  
1093 fluctuating between the daily minimum and maximum temperature (dotted line) and  
1094 the number (L3 kg DM-1) observed by Wilkie et al. (submitted; points and error bars  
1095 show mean and 95% confidence interval). Predicted numbers of L3 on herbage  
1096 using a single mortality rate for L3 on pasture and no vertical migration (dotted line;  
1097 Smith, 1990) are superimposed for comparison.



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1100 Figure 7. The number of L3 on pasture (soil and herbage combined) predicted for  
1101 *Haemonchus contortus* (top panel), *Teladorsagia circumcincta* (middle panel) and  
1102 *Ostertagia ostertagi* (bottom panel) when using historical climatic data for the period  
1103 1969-1999 (solid line) and climatic data based on the RCP8.5 high emissions climate  
1104 change scenario for the period 2070-2100 (broken line). Data shown are the  
1105 disaggregated annual means from the thirty year time-series. The first year of each  
1106 time series was discarded. One hundred new eggs were input daily. Therefore no  
1107 assumptions were made regarding management or intensity of infection in the host  
1108 and the predicted dynamics are entirely climate-driven.



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1112 Table 1. State variable and parameter definitions

<b>State variable/ Parameter</b>	<b>Definition</b>	<b>Units</b>
<b><i>E</i></b>	Eggs	-
<b><i>L</i></b>	First stage (L1) and second stage (L2) larvae	-
<b><i>L3<sub>f</sub></i></b>	Third stage infective larvae (L3) in faeces	-
<b><i>L3<sub>p</sub></i></b>	Total L3 on pasture (soil and herbage combined)	-
<b><i>L3<sub>s</sub></i></b>	L3 in soil	-
<b><i>L3<sub>h</sub></i></b>	L3 on herbage	-
<b><i>σ</i></b>	Development rate from egg to L3	Instantaneous daily rate
<b><i>μ<sub>1</sub></i></b>	Egg mortality rate	Instantaneous daily rate
<b><i>μ<sub>2</sub></i></b>	L1 and L2 mortality rate	Instantaneous daily rate
<b><i>μ<sub>3</sub></i></b>	L3 mortality rate in faeces	Instantaneous daily rate
<b><i>μ<sub>4</sub></i></b>	L3 mortality rate in soil	Instantaneous daily rate
<b><i>μ<sub>5</sub></i></b>	L3 mortality rate on herbage	Instantaneous daily rate
<b><i>m<sub>1</sub></i></b>	Horizontal migration (translation) of L3 onto pasture	Instantaneous daily rate
<b><i>m<sub>2</sub></i></b>	Proportion of total pasture L3 on herbage	Proportion
<b><i>C</i></b>	Development success correction factor	Proportion

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1115 Table 2. Parameter estimates derived from data in the literature and additional  
 1116 laboratory experiments. ANOVA results are shown for the linear models fitted to  
 1117 data from the literature to estimate temperature-dependent rates (see text).

Parameter	Species <sup>a</sup>	Estimate <sup>b</sup>	Data source
$\delta$	<i>Hc</i>	$-0.09746 + 0.01063T$ ( $F_{1,6}=43.5$ , $p<0.001$ , $R^2=0.88$ , $R^2_{adj}=0.86$ )	Hsu and Levine, 1977; Rose, 1963
	<i>Tc</i>	$-0.02085 + 0.00467T$ ( $F_{1,10}=76.57$ , $p<0.001$ , $R^2=0.88$ , $R^2_{adj}=0.87$ )	Crofton and Whitlock, 1965; Crofton, 1965; Pandey et al., 1989; Salih and Grainger, 1982; Young et al., 1980a
	<i>Oo</i>	$-0.07258 + 0.00976T$ ( $F_{1,8}=76.14$ , $p<0.001$ , $R^2=0.90$ , $R^2_{adj}=0.89$ )	Pandey, 1972a; Rose, 1961; Young et al., 1980b
$\mu_1$	<i>Hc</i>	$\exp(-1.47135 - 0.11444T$ $+ 0.00327T^2)$ ( $F_{2,3}=4.65$ , $p=0.12$ , $R^2=0.76$ , $R^2_{adj}=0.59$ )	Todd et al., 1976a
	<i>Tc</i>	$\exp(-1.62026 - 0.17771T$ $+ 0.00629T^2)$ ( $F_{2,2}=6.27$ , $p=0.27$ , $R^2=0.93$ , $R^2_{adj}=0.78$ )	Pandey et al., 1993, 1989
	<i>Oo</i>	$\exp(-4.38278 - 0.10640T$ $+ 0.00540T^2)$ ( $F_{2,1}=6.27$ , $p=0.06$ , $R^2=0.99$ , $R^2_{adj}=0.99$ )	Pandey, 1972
$\mu_2$	<i>Hc</i>	$\exp(-1.82300 - 0.14180T$ $+ 0.00405T^2)$ ( $F_{2,1}=1.723^{31}$ , $p<0.001$ , $R^2=1$ , $R^2_{adj}=1$ ) <sup>c</sup>	Todd et al., 1976a
	<i>Tc, Oo</i>	As $\mu_1$ above	-
$\mu_3$	<i>Hc</i>	$\exp(-2.63080 - 0.14407T$ $+ 0.00463T^2)$ ( $F_{2,9}=8.48$ , $p=0.008$ , $R^2=0.65$ , $R^2_{adj}=0.58$ )	Todd et al., 1976a, 1976b

	$Tc, Oo$	$10 * \mu_4$	Pandey, 1972; Persson, 1974a
$\mu_4$	$Hc$	$\exp(-3.68423 - 0.25346T + 0.00740T^2)$ ( $F_{2,8}=50.76, p<0.001, R^2=0.93, R^2_{adj}=0.91$ )	Jehan and Gupta, 1974; Todd et al., 1976b
	$Tc$	$\exp(-4.58817 - 0.13996T + 0.00461T^2)$ ( $F_{2,12}=43.55, p<0.001, R^2=0.88, R^2_{adj}=0.86$ )	Gruner and Suryahadi, 1993; Jasmer et al., 1987; Pandey et al., 1993; Rossanigo and Gruner, 1996
	$Oo$	$\exp(-6.388 - 0.2681T + 0.01633T^2 - 0.00016T^3)$ ( $F_{3,5}=28.81, p=0.001, R^2=0.95, R^2_{adj}=0.91$ )	Pandey, 1972
$\mu_5$	$Hc, Tc, Oo$	As $\mu_3$ above	Grenfell et al., 1986; van Dijk and Morgan, 2011; van Dijk et al., 2009
$m_1$	$Hc$	$\begin{cases} 0.25, & P \geq 2 \\ 0, & P < 2 \text{ AND } \sum_{i=-4}^t P_i/E_i < 1 \\ 0.051, & P < 2 \text{ AND } \sum_{i=-4}^t P_i/E_i \geq 1 \end{cases}$	Present study; O'Connor et al., 2008; Wang et al., 2014
	$Tc$	$\begin{cases} 0.21, & P \geq 2 \\ 0, & P < 2 \text{ AND } \sum_{i=-7}^t P_i/E_i < 1 \\ 0.025, & P < 2 \text{ AND } \sum_{i=-7}^t P_i/E_i \geq 1 \end{cases}$	Present study; O'Connor et al., 2008; Wang et al., 2014
	$Oo$	$\begin{cases} 0.06, & P \geq 2 \\ 0, & P < 2 \end{cases}$	Grønvold and Høgh-Schmidt, 1989



$m_2$	$Hc, Tc,$ $Oo$	$\exp(-5.48240 + 0.45392T - 0.01252T^2)$ ( $F_{2,1}=442.9, p=0.034, R^2>0.99,$ $R^2_{adj}>0.99$ )	Callinan and Westcott, 1986
<b>C</b>	$Hc$	$\begin{cases} 0.1, & \sum_{i=4}^t P_i/E_i < 1 \\ 1, & \sum_{i=4}^t P_i/E_i \geq 1 \end{cases}$	
	$Tc$	$\begin{cases} 0.1, & \sum_{i=7}^t P_i/E_i < 1 \\ 1, & \sum_{i=7}^t P_i/E_i \geq 1 \end{cases}$	

1118 <sup>a</sup>  $Hc = Haemonchus contortus, Tc = Teladorsagia circumcincta, Oo = Ostertagia$   
1119  $ostertagi$

1120 <sup>b</sup>  $T = \text{temperature } (^{\circ}\text{C}), P = \text{total daily precipitation (mm)}, E = \text{total daily}$   
1121  $\text{evapotranspiration (mm)}$

1122 <sup>c</sup> Note that the statistical significance here is an artefact of overfitting

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1125 Table 3. Validation of simulations using data provided in the literature. Models are  
1126 considered a good fit if regression through the origin is significant and the slope is  
1127 not significantly different from 1. The error and  $R^2$  are used to compare competing  
1128 models.

Data source	Species <sup>a</sup>	Model component tested	Temperature data used	Error (residual sum of squares)	Linear regression	R <sup>2</sup> (R <sup>2</sup> <sub>adjusted</sub> )	Slope (95% CI <sup>b</sup> )
Rose, 1963	<i>Hc</i>	Faeces (eq. 1-3; D50)	Min - Max	1093.5	F <sub>1,6</sub> = 12.21, p=0.013	0.67 (0.62)	2.18 (0.93 – 3.43)
Rose, 1963	<i>Hc</i>	Faeces (eq. 1-3; D50)	Mean	1097.5	F <sub>1,6</sub> = 9.561, p=0.021	0.61 (0.55)	1.93 (0.68 – 3.17)
Rose, 1963	<i>Hc</i>	Faeces (eq. 1-3; development success)	Min - Max	31.28	F <sub>1,11</sub> = 28.53, p<0.001	0.72 (0.70)	1.28 (0.80 – 1.76)
Rose, 1963	<i>Hc</i>	Faeces (eq. 1-3; development success)	Mean	105.74	F <sub>1,11</sub> = 11.91, p=0.005	0.52 (0.48)	0.49 (0.20 – 0.77)
Wilkie et al. submitted	<i>Hc</i>	Pasture (eq. 4)	Min - Max	1.83 x 10 <sup>8</sup>	F <sub>1,3</sub> = 6.78, p=0.080	0.69 (0.59)	0.94 (0.22 – 1.67)
Wilkie et al. submitted	<i>Hc</i>	Pasture (eq. 4)	Mean	1.43 x 10 <sup>8</sup>	F <sub>1,3</sub> = 16.72, p=0.026	0.85 (0.80)	1.48 (0.76 – 2.20)
Rossani go and Gruner, 1995	<i>Tc</i>	Faeces (eq. 1-3; development success)	Constant	1679.10	F <sub>1,9</sub> = 86.9, p<0.001	0.91 (0.90)	1.54 (1.21 – 1.86)
Wilkie et al. submitted	<i>Tc</i>	Pasture (eq. 4)	Min - Max	1.03 x 10 <sup>10</sup>	F <sub>1,3</sub> = 9.91, p=0.051	0.77 (0.69)	1.26 (0.46 – 2.07)

Wilkie et al. submitted	<i>Tc</i>	Pasture (eq. 4)	Mean	1.17 x 10 <sup>10</sup>	F <sub>1,3</sub> = 17.84, p=0.024	0.86 (0.81)	1.75 (0.92-2.58)
Rossanigo and Gruner, 1995	<i>Oo</i>	Faeces (eq. 1-3; development success)	Constant	2050.2	F <sub>1,5</sub> = 36.86, p=0.002	0.88 (0.86)	0.84 (0.57-1.12)
Rose, 1961	<i>Oo</i>	Faeces (eq. 1-3; D50)	Min – Max		F <sub>1,11</sub> = 26.9, p<0.001	0.71 (0.68)	21.03 (12.9 – 29.1)
Rose, 1961	<i>Oo</i>	Faeces (eq. 1-3; D50)	Mean		F <sub>1,11</sub> = 98.51, p<0.001	0.90 (0.89)	0.62 (0.50 – 0.75)
Wilkie et al. submitted	<i>Hc</i>	Smith (1990)	Mean	2.66 x 10 <sup>9</sup>	F <sub>1,3</sub> = 4.73, p=0.12	0.61 (0.48)	0.28 (0.02-0.53)

1129 <sup>a</sup> *Hc* = *Haemonchus contortus*, *Tc* = *Teladorsagia circumcincta*, *Oo* = *Ostertagia*  
1130 *ostertagi*

1131 <sup>b</sup> 95% confidence intervals were estimated as 2 x the standard error of the slope  
1132 coefficient.

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